

March 30, 2004

**Cayuga Lake Water Quality Monitoring,
Related to the LSC Facility: 2003**

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1. Objective/Study Area

The primary objective is to conduct an ambient water quality monitoring program focusing on the southern portion of Cayuga Lake to support long-term records of trophic state indicators, including concentrations of phosphorus and chlorophyll, and Secchi disc transparency, and other measures of water quality.

Cayuga Lake is the second largest of the Finger Lakes. A comprehensive limnological description of the lake has been presented by Oglesby (1979). The lake is monomictic (stratifies in summer), mesotrophic (intermediate level of biological productivity), and is a hardwater alkaline system. Much of the tributary inflow received by the lake enters at the southern end; e.g., ~ 40% is contributed by the combination of Fall Creek and Cayuga Inlet (Figure 1). Effluent from two domestic wastewater treatment (WWT) facilities also enters this portion of the lake (Figure 1a). The discharge from Cornell's LSC facility enters the southern portion (e.g., south of McKinney's Point) of the lake along the east shore (Figure 1a). The LSC facility started operating in early July of 2000.

2. Design

2.1. Description of Parameters Selected for Monitoring

2.1.1. Phosphorus (P)

Phosphorus (P) plays a critical role in supporting plant growth. Phosphorus has long been recognized as the most critical nutrient controlling phytoplankton (microscopic plants of the open waters) growth in most lakes in the north temperate zone. Degradation in water quality has been widely documented for lakes that have received excessively high inputs of P from man's activities. Increases in P inputs often cause increased growth of phytoplankton in lakes. Occurrences of particularly high concentrations of phytoplankton are described as "blooms". The accelerated "aging" of lakes associated with inputs of P from man's activities has been described as cultural eutrophication.

The three forms of P measured in this monitoring program, total P (TP), total dissolved P (TDP), and soluble reactive P (SRP), are routinely measured in many limnological and water quality programs. TP is widely used as an indicator of trophic state (level of plant production). TDP and SRP are measured on filtered (0.45 μm) samples. Most TDP is assumed to be ultimately available to support phytoplankton growth. SRP is a component of TDP that is usually assumed to be immediately available to support phytoplankton growth. Particulate P (PP; incorporated in, or attached to, particles) is calculated as the difference between paired measurements of TP and TDP. The composition of PP can vary greatly in time for a particular lake, and between different lakes. Contributing components include phytoplankton and other P-bearing particles that may be resuspended from the bottom or received from stream/river inputs.

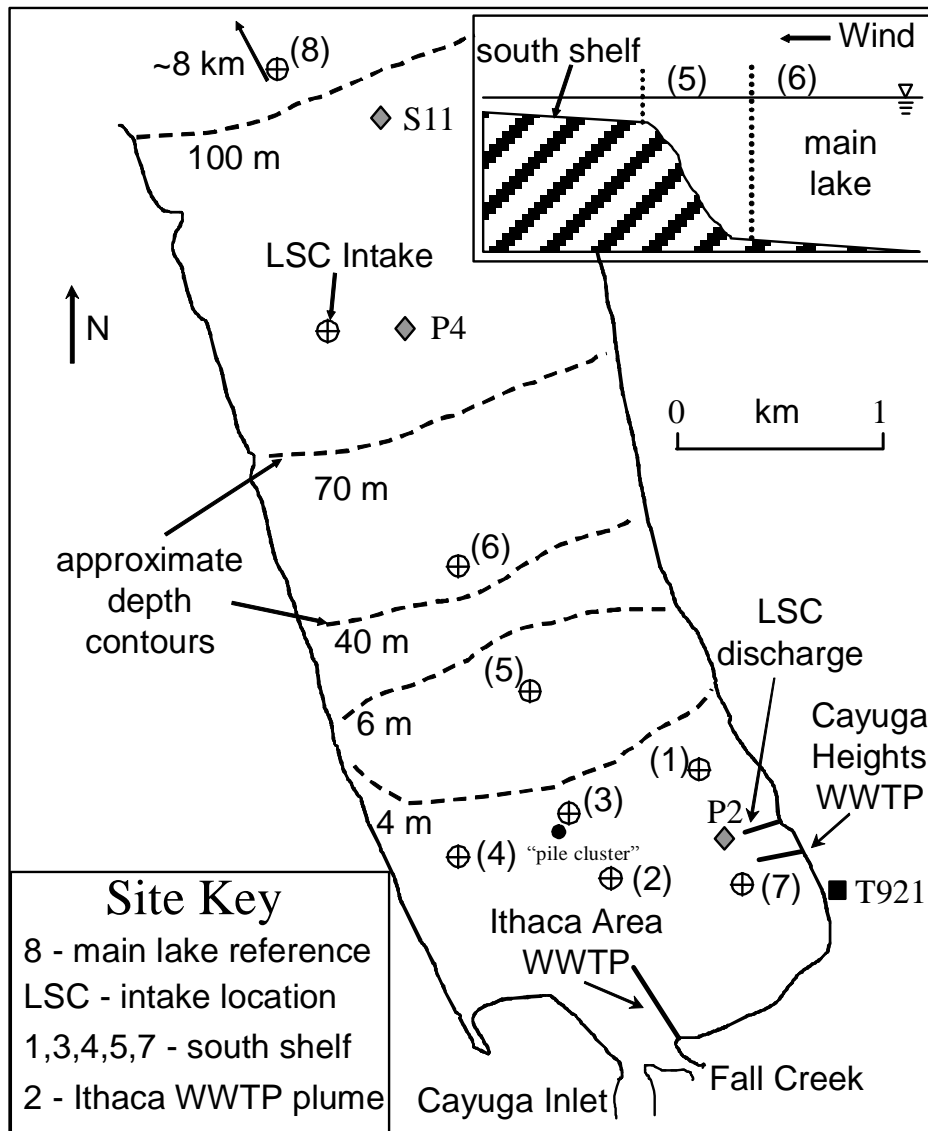


Figure 1a. Sampling sites, setting, approximate bathymetry, for LSC monitoring program, southern end of Cayuga Lake. Sites sampled during 1994 – 1996 study (P2, P4 and S11; Stearns and Wheler 1997) are included for reference. Locations of sampling sites and outfalls are approximate.

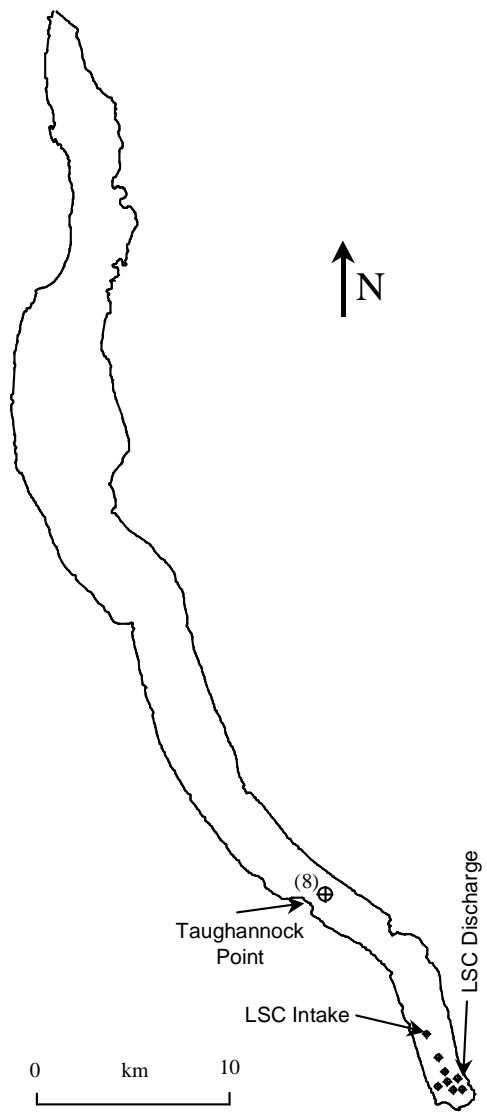


Figure 1b. Sampling sites for LSC monitoring program, within the context of the entire Cayuga Lake basin.

2.1.2. Clarity/Optical Properties

The extent of the penetration of light in water (e.g., ability to see submerged objects), described as clarity, is closely coupled to the public's perception of water quality. Light penetration is particularly sensitive to the concentration, composition and size of particles. In lakes where phytoplankton are the dominant component of the particle population, measures of clarity may be closely correlated to concentrations of TP and phytoplankton biomass (e.g., as measured by chlorophyll). Clarity is relatively insensitive to phytoplankton biomass when and where concentrations of other types of particles are high. In general, light penetration is low when concentrations of phytoplankton, or other particles, are high.

Two measures of light penetration are made routinely in this program, Secchi disc transparency (in the field) and turbidity (laboratory). The Secchi disc measurement has a particularly long history in limnological studies, and has proven to be a rather powerful piece of information, even within the context of modern optical measurements. It remains the most broadly used measure of light penetration. The higher the Secchi disc measurement the greater the extent of light penetration. Turbidity, as measured with a nephelometric turbidimeter, measures the light captured from a standardized source after passage through a water sample. Turbidity and Secchi disc depth are regulated by a heterogeneous population of suspended particles that include not only phytoplankton, but also clay, silt, and other finely divided organic and inorganic matter. The higher the turbidity value the higher the concentration of particles that limit light penetration.

Two other optical measurements are made as part of this program, irradiance and beam attenuation. These parameters are included to augment the information concerning light penetration. Depth profiles of irradiance are collected to determine the attenuation (or extinction) coefficient, another measure of light penetration.

2.1.3. Chlorophyll/Fluorescence

Chlorophyll **a** is the principal photosynthetic pigment that is common to all phytoplankton. Chlorophyll (usually as chlorophyll **a**) is the most widely used surrogate measure of phytoplankton biomass, and is generally considered to be the most direct and reliable measure of trophic state. Increases in chlorophyll concentrations indicate increased phytoplankton production. The major advantages of chlorophyll as a measure of phytoplankton biomass are: (1) the measurement is relatively simple and direct, (2) it integrates different types and ages of phytoplankton, (3) it accounts to some extent for viability of the phytoplankton, and (4) it is quantitatively coupled to optical properties that may influence clarity. However, the chlorophyll measurement does not resolve phytoplankton type, and the chlorophyll content per unit biomass can vary according to species and ambient environmental conditions. Therefore, it is an imperfect measure of phytoplankton biomass. Fluorescence has been widely used as a surrogate measure of chlorophyll. Fluorescence measurements are made in the field in this program.

Rather wide variations in chlorophyll concentrations can occur seasonally, particularly in productive lakes. The details of the timing of these variations, including the occurrence of

blooms, often differ year-to-year. Seasonal changes in phytoplankton biomass reflect imbalance between growth and loss processes. Factors influencing growth include nutrient availability (concentrations), temperature and light. Phytoplankton are removed from the lake either by settling, consumption by small animals (e.g., zooplankton), natural death, or exiting the basin. During intervals of increases in phytoplankton, the rate of growth exceeds the summed rates of the various loss processes.

2.1.4. Temperature

Temperature is a primary regulator of important physical, chemical, and biochemical processes in lakes. It is perhaps the most fundamental parameter in lake monitoring programs. Lakes in the northeast go through major temperature transformations linked primarily to changes in air temperature and incident light. Important cycles in aquatic life and biochemical processes are linked to the annual temperature cycle. Deep lakes stratify in summer in this region, with the warmer less dense water in the upper layers (epilimnion) and the colder more dense water in the lower layers (hypolimnion). A rather strong temperature/density gradient in intermediate depths between the epilimnion and hypolimnion (metalimnion) limits cycling of materials from the hypolimnion to the epilimnion during summer. Gradients in temperature are largely absent over the late fall to spring interval, allowing active mixing throughout the watercolumn (e.g., turnover).

2.2. Timing

Lake sampling and field measurements were conducted by boat during the spring to fall interval of 2003, beginning in April and extending through October. The full suite of laboratory and field measurements was made for 16 bi-weekly monitoring trips. Additionally, recording thermistors were deployed continuously at one location; temperature measurements were made hourly over the April – October interval. The thermistors were exchanged biweekly with fresh units for data downloading and maintenance. Deployments made on October 29, 2003 will be retrieved in April 2004. Measurements are recorded on a daily basis over this later interval. Laboratory measurements of phosphorus concentration, T_n , dissolved oxygen concentration (DO), and pH were made on samples from the LSC influent and effluent collected weekly during operation of the LSC facility.

2.3. Locations

An array of sampling sites (e.g., grid) has been adopted that provides a robust representation of the southern portion of the lake (Figure 1a and b). This sampling grid may reasonably be expected to resolve persistent water quality gradients that may be imparted by the various inputs/inflows that enter this portion of the lake. Further, inclusion of these sites is expected to contribute to fair representation of average conditions for this portion of the lake.

Seven sites were monitored for the full suite of parameters in the southern end of the lake (sites 1 through 7). The intake location for the LSC facility and site 8, located further north as a reference for the main lake conditions, were also sampled. Positions (latitude, longitude) for the eight sites are specified in Table 1. The configuration of sites includes two transect lines; one

with 3 sites along an east-west line extending from an area near the discharge location, the other with 4 sites running approximately along the main axis of the lake (Figure 1a). Additionally, two sites (1 and 7) bound the location of the LSC discharge, paralleling the east shore (Figure 1a). The position for thermistor deployment (“pile cluster”) is shown in Figure 1 and specified in Table 1. The “Global Positioning System” (GPS) was used to locate the sampling/monitoring sites. A reference position located at the southern end of the lake (T921; Figure 1a) was used to assess the accuracy of the GPS for each monitoring trip.

Table 1: Specification of site locations (GPS) and depths (sonar) for ambient water quality monitoring (refer to Figure 1a). Sites sampled during 1994 – 1996 study (P2, P4 and S11; Stearns and Wheler 1997) are included for reference.

Site No.	Latitude	Longitude	Depth (m)
1 (discharge boundary)	42°28.3'	76°30.5'	5
2	28.0'	30.8'	3
3	28.2'	30.9'	4
4	28.2'	31.4'	4
5	28.5'	31.1'	6
6	28.8'	31.3'	40
7 (discharge boundary)	28.0'	30.3'	3.5
8 (off Taughannock Pt.)	33.0'	35.0'	110
thermistor “pile cluster”	28.1'	31.0'	4
LSC Intake	29.4'	31.8'	78
P2	28.20'	30.40'	4
P4	29.31'	31.41'	65
S11	29.60'	31.45'	72

2.4. Field Measurements/Seabird Profiling

Instrumentation profiles were collected in the field at 9 locations (sites 1 through 8 and the LSC Intake; Figure 1a) with a SeaBird profiler. Profiles extended from the surface to within 2m of the lake bottom, or to 20 m at deeper sites. Deeper profiles were obtained for the intake site. Parameters measured in the profiles and the potential utility of the information are summarized in Table 2. Additionally, dissolved oxygen was measured at site 3 each monitoring trip with a HydroLab Surveyor 3, calibrated and operated according to the manufacturer’s specifications. Secchi disc transparency was measured at all sites with a 20 cm diameter black and white quadrant disc (Wetzel and Likens 1991).

Table 2: SeaBird profiler: parameters and utility.

Parameter	Utility
Temperature	heat budget, density stratification
Conductivity	tracer, mixing patterns
Fluorescence	measure of chlorophyll
Beam attenuation	identification of particle rich layers, including benthic nepheloid layers
Irradiance	determination of attenuation
Scalar	coefficients
Downwelling	

2.5. Field Methods

Water samples were collected with a well-rinsed Van Dorn sampler or submersible pump, with depths marked on the line/hose. Care was taken that the sampling device was deployed vertically within the water column at the time of sampling. Samples for laboratory analysis were composite-type, formed from equal volumes of sub-samples collected at depths of 0, 2 and 4 meters for sites 5, 6, LSC, and 8. Composite samples for sites 1, 2, 3, 4, and 7 were formed from equal volumes of sub-samples collected at depths of 0 and 2 meters. The composite-type samples avoid over-representation of the effects of temporary secondary stratification in monitored parameters. In addition, samples were collected at the LSC intake site at 1m and 3m above the bottom (depth of ~ 77m). Sample bottles were stored in ice and transported to the laboratory on the same day of sampling. Chain of custody procedures were observed for all samples collected for laboratory analysis.

2.6. Laboratory Analyses, Protocols

Laboratory analyses for the selected parameters were conducted according to methods specified in Table 3. Detection limits for these analyses are also included. Most of these laboratory analyses are “Standard Methods”. Results below the limit of detection are reported as ½ the limit of detection. The chlorophyll methods are among the most commonly used in lake studies. Chlorophyll concentrations were determined by fluorometric assay (USEPA 1992) for samples collected on 4/15/03, 4/24/03, 5/8/03, 6/19/03 and 7/2/03, and by spectrophotometric analysis (Parsons et al. 1984) for all other samples. The fluorometric concentrations were adjusted to achieve equivalency with spectrophotometric determinations according to the regression expression developed from paired assays on samples collected from Muscoot and New Croton Reservoirs, New York (UFI, unpublished data):

$$\text{Chl } a_{\text{SPEC}} = 1.21 \times \text{Chl } a_{\text{FLUOR}} + 0.24; r^2 = 0.95$$

The acidified turbidity method has been applied by this study team for a number of hard water systems such as Cayuga Lake. Specifications adhered to for processing and preservation of

samples, containers for samples, and maximum holding times before analyses, are summarized in Table 4.

2.7. Quality Assurance/Control Program

A quality assurance/control (QA/QC) program was conducted to assure that ambient lake data collected met data quality objectives for precision, accuracy, representativeness, comparability, and completeness.

Table 3: Specification of laboratory methods for ambient water quality monitoring.

Analyte	Method No.	Reference	Limit of Detection
total phosphorus	4500-P	APHA (1992)	0.6 $\mu\text{g}\cdot\text{L}^{-1}$
soluble reactive phosphorus	4500-P	APHA (1992)	0.3 $\mu\text{g}\cdot\text{L}^{-1}$
total dissolved phosphorus	4500-P	APHA (1992)	0.6 $\mu\text{g}\cdot\text{L}^{-1}$
turbidity	2130-B	APHA (1992)	-
acidified turbidity		Effler and Johnson (1987)	-
chlorophyll a		Parsons et al. (1984)	0.4 $\mu\text{g}\cdot\text{L}^{-1}$
	445.0	USEPA (1992)	0.4 $\mu\text{g}\cdot\text{L}^{-1}$

2.7.1. Field Program

Precision of sampling and sample handling was assessed by a program of field replicates. Samples for laboratory analyses were collected in triplicate at site 1 on each sampling day. Triplicate samples were collected at one of the other eight stations each monitoring trip. This station was rotated each sampling trip through the field season. Secchi disc measurements were made in triplicate at all sites through the field season. Precision was high for the triplicate sampling/measurement program, as represented by the average values of the coefficient of variation for the 2003 program (Table 5).

2.7.2. Laboratory Program

The laboratory quality assurance/control program conducted was as specified by the National Environmental Laboratory Accreditation Program (NELAP 2003). NELAP methods were used to assure precision and accuracy, completeness and comparability (NELAP 2003). The program included analyses of reference samples, matrix spikes, blind proficiency samples, and duplicate analyses. Calibration and performance evaluation of analytical methods was as specified in the NELAP program; this includes control charts of reference samples, matrix spikes, and duplicate analyses.

Table 4: Summary of processing, preservation, storage containers and holding times for laboratory measurements; see codes below.

Parameter	Processing	Preservation	Container	Holding Time
total phosphorus	c	a	1	1
soluble reactive phosphorus	a	b	1	2
total dissolved phosphorus	a	a	1	1
chlorophyll a	b	c	2	3
turbidity	c	b	2	2
acidified turbidity	d	b	2	2

codes for Table 4:

processing: a - filter with 0.45 µm cellulose acetate filter
 b - filter with 0.45 µm cellulose nitrate filter
 c - whole water sample
 d - acidified to pH = 4.3 for 1 min.

preservation: a - H₂SO₄ to pH < 2
 b - none
 c - store filter frozen until analysis

container: 1 - 250 ml acid washed borosilicate boston round
 2 - 4L polypropylene container

holding time: 1 - 28 days
 2 - 24 hours
 3 - 200 days

Table 5: Precision for triplicate sampling/measurement program for key parameters for 2003, represented by the average coefficient of variation.

Parameter	Site 1	Rotating Site*
total phosphorus	0.11	0.11
chlorophyll a	0.10	0.14
turbidity	0.11	0.13
Secchi disc	< 0.01	< 0.01

* average of Sites 2, 3, 4, 5, 6, 7, 8, LSC

3. Results, 2003

The measurements made in the 2003 monitoring program are presented in two formats here: (1) in tabular form (Table 6) as selected summary statistics for each site, and (2) as time plots (Figure 2) for selected sites and site groupings. Detailed listings of data are presented in Appendix I. LSC Discharge Monitoring Report Data are presented in Appendix 2. The adopted summary statistics include the mean, the range of observations, and the coefficient of variation ($CV = \text{standard deviation}/\text{mean}$; Table 6). The plots present three time series; these include (except for Secchi disc) one for site 2, another for site 8, and the third is an “average” of sites intended to represent overall conditions in the southern portion of the lake. This southern portion is designated as the “shelf”, as depths are less than 6 m. The “average” for the shelf is the mean of observations for sites 3, 4, 5, and the average of sites 1 and 7 (together to represent conditions in the eastern portion of the study area; see Figure 1a). Observations for site 6 are not included in this averaging because this location, while proximate, is in deeper water (> 40 m; i.e., off the shelf). Measurements at site 8 are presented separately in these plots to reflect lake-wide (or the main lake) conditions. Observations for site 2 are separated from the other sites of the southern end because the results indicate this location is at times within the discharge plume of the Ithaca Area WWTP. Time series for site 2 appear as insets in the time plots (Figure 2) to accommodate the greater magnitudes of some of the observations for this site, and still allow resolution of temporal structure observed for other locations. The Secchi disc plot (Figure 2e) presents observations for sites 6, LSC, and 8; the deeper sites, where observations were always less than the bottom depth. Time series for the LSC influent, the LSC effluent, and the shelf are presented separately (Figure 2j-l). Paired profiles of temperature, the beam attenuation coefficient (BAC), and chlorophyll fluorescence obtained at the LSC intake site on 15 monitoring dates in 2003 are also presented (Figure 3).

Previous annual reports (UFI 1999, 2000, 2001, 2002, 2003) documented occurrences of extremely high concentrations of forms of phosphorus (TP, TDP, and SRP) and nitrogen (TDN and T-NH₃) at site 2. These occurrences are likely associated with the proximity of site 2 to the Ithaca Area WWTP discharge (Figure 1a) enriched in these components. High concentrations of phosphorus continued to be observed at this site in 2003 (Figure 2a-c). Site 2 is omitted in the formation of the average for the shelf because the effect is localized, temporally irregular, and is representative of only a relatively small volume of water.

Table 6: Summary of results of monitoring program according to site, 2003.

TP ($\mu\text{gP}\cdot\text{L}^{-1}$)			
SITE	MEAN	CV	RANGE
1	18.5	0.52	9.5 – 42.8
2	35.2	0.52	13.2 – 71.6
3	18.1	0.36	8.8 – 30.8
4	15.9	0.64	7.0 – 47.6
5	14.0	0.33	8.8 – 23.1
6	12.2	0.23	7.5 – 18.2
7	28.4	0.93	10.3 – 112.3
8	10.5	0.19	7.2 – 16.3
LSC	10.9	0.25	7.5 – 17.6

Chl a ($\mu\text{g}\cdot\text{L}^{-1}$)			
SITE	MEAN	CV	RANGE
1	4.5	0.52	1.1 – 9.5
2	5.6	0.63	1.7 – 12.9
3	5.4	0.67	1.7 – 13.1
4	3.8	0.83	1.5 – 14.6
5	4.7	0.64	1.7 – 14.0
6	5.0	0.58	1.9 – 14.2
7	6.5	0.76	1.7 – 17.7
8	4.6	0.50	1.4 – 11.0
LSC	4.6	0.59	2.0 – 13.2

TDP ($\mu\text{gP}\cdot\text{L}^{-1}$)			
SITE	MEAN	CV	RANGE
1	5.6	0.77	1.8 – 17.4
2	8.4	0.80	1.6 – 24.2
3	4.8	0.66	1.6 – 11.1
4	4.2	0.66	1.3 – 11.8
5	4.2	0.63	1.2 – 9.6
6	3.8	0.59	1.6 – 9.4
7	7.1	1.04	1.3 – 32.6
8	3.4	0.57	0.7 – 7.5
LSC	3.7	0.55	1.0 – 8.1

T_n (NTU)			
SITE	MEAN	CV	RANGE
1	2.4	0.83	1.0 – 7.6
2	5.4	1.04	0.9 – 20.0
3	2.4	0.75	0.6 – 8.1
4	2.1	1.21	0.5 – 10.9
5	1.8	0.61	0.7 – 3.7
6	1.3	0.57	0.5 – 2.8
7	4.8	1.59	1.0 – 31.1
8	1.1	0.41	0.6 – 2.1
LSC	1.1	0.44	0.6 – 2.2

SRP ($\mu\text{gP}\cdot\text{L}^{-1}$)			
SITE	MEAN	CV	RANGE
1	2.6	1.32	0.2 – 11.9
2	4.6	1.09	0.2 – 15.0
3	2.0	1.19	0.2 – 6.0
4	1.6	1.32	0.2 – 6.3
5	1.7	1.31	0.2 – 6.0
6	1.3	1.52	0.2 – 5.7
7	3.1	1.76	0.2 – 21.9
8	1.2	1.26	0.2 – 4.4
LSC	1.3	1.46	0.2 – 5.6

Temperature ($^{\circ}\text{C}$) @ 2m			
SITE	MEAN	CV	RANGE
1	14.1	0.48	4.0 – 26.3
2	14.1	0.45	5.9 – 25.1
3	14.0	0.49	3.9 – 26.6
4	14.2	0.47	5.4 – 26.7
5	14.2	0.49	4.3 – 26.4
6	14.2	0.49	4.3 – 26.4
7	14.1	0.46	4.8 – 25.8
8	14.2	0.53	3.0 – 26.0
LSC	14.1	0.52	3.2 – 26.2

Table 6 (cont.): Summary of results of monitoring program according to site, 2003.

Beam Attenuation Coeff. (m^{-1}) @ 2m			
SITE	MEAN	CV	RANGE
1	2.1	0.56	0.9 – 4.8
2	4.4	0.97	1.0 – 15.9
3	2.5	0.80	0.8 – 9.1
4	2.0	1.05	0.6 – 9.4
5	1.7	0.40	0.7 – 2.7
6	1.8	0.94	0.7 – 7.6
7	2.6	0.62	0.9 – 7.1
8	1.3	0.46	0.6 – 2.4
LSC	1.3	0.45	0.7 – 2.5

K_s Attenuation Coeff. (m^{-1})			
SITE	MEAN	CV	RANGE
1	-	-	-
2	-	-	-
3	-	-	-
4	-	-	-
5	-	-	-
6	0.32	0.22	0.26 – 0.52
7	-	-	-
8	0.31	0.20	0.20 – 0.45
LSC	0.30	0.20	0.24 – 0.42

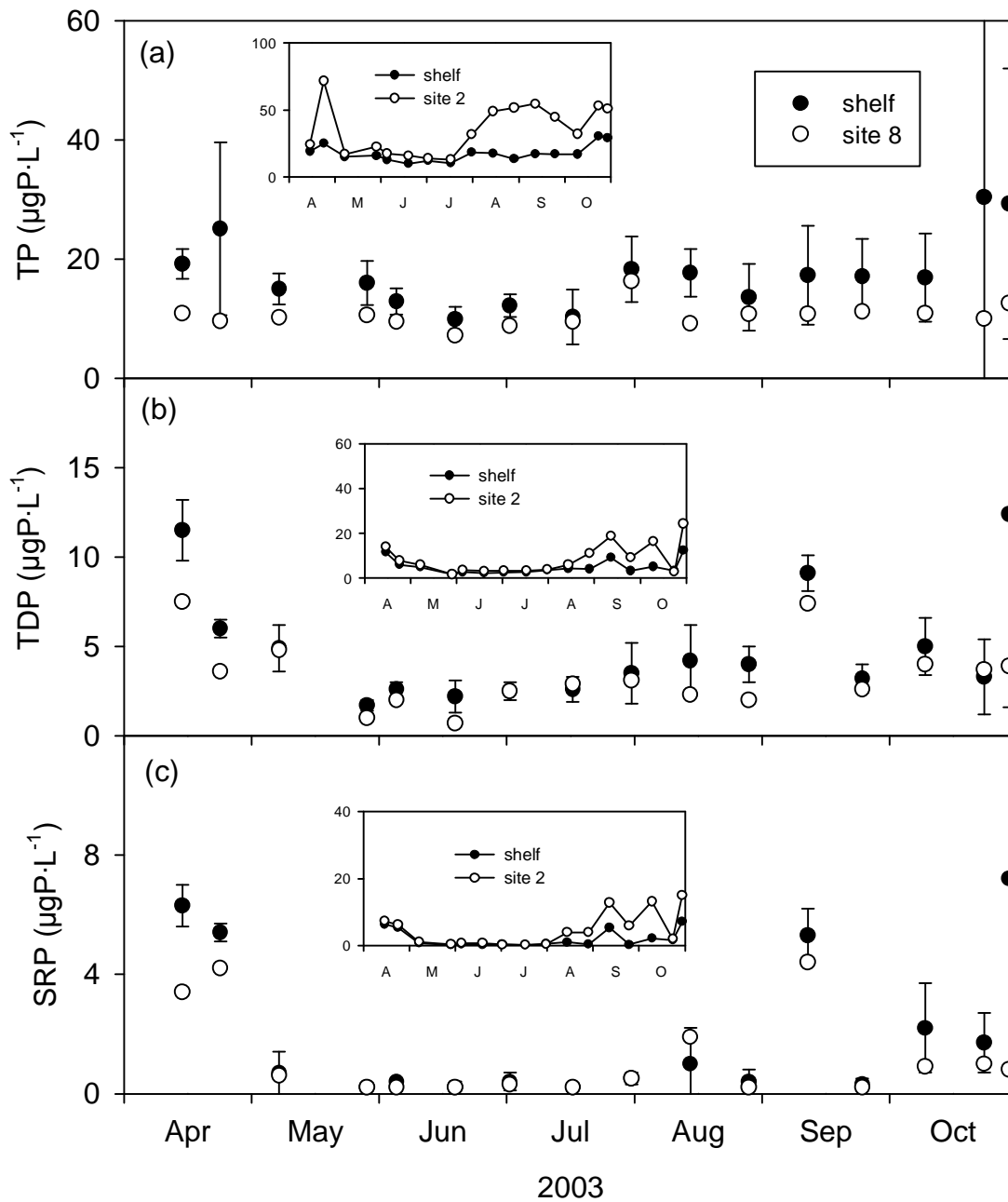


Figure 2a-c. Time series of parameter values for Cayuga Lake for 2003: (a) TP, (b) TDP, and (c) SRP. Insets present results for site 2. Results for the “shelf” are averages; the error bars represent spatial variation with dimensions of ± 1 standard deviation.

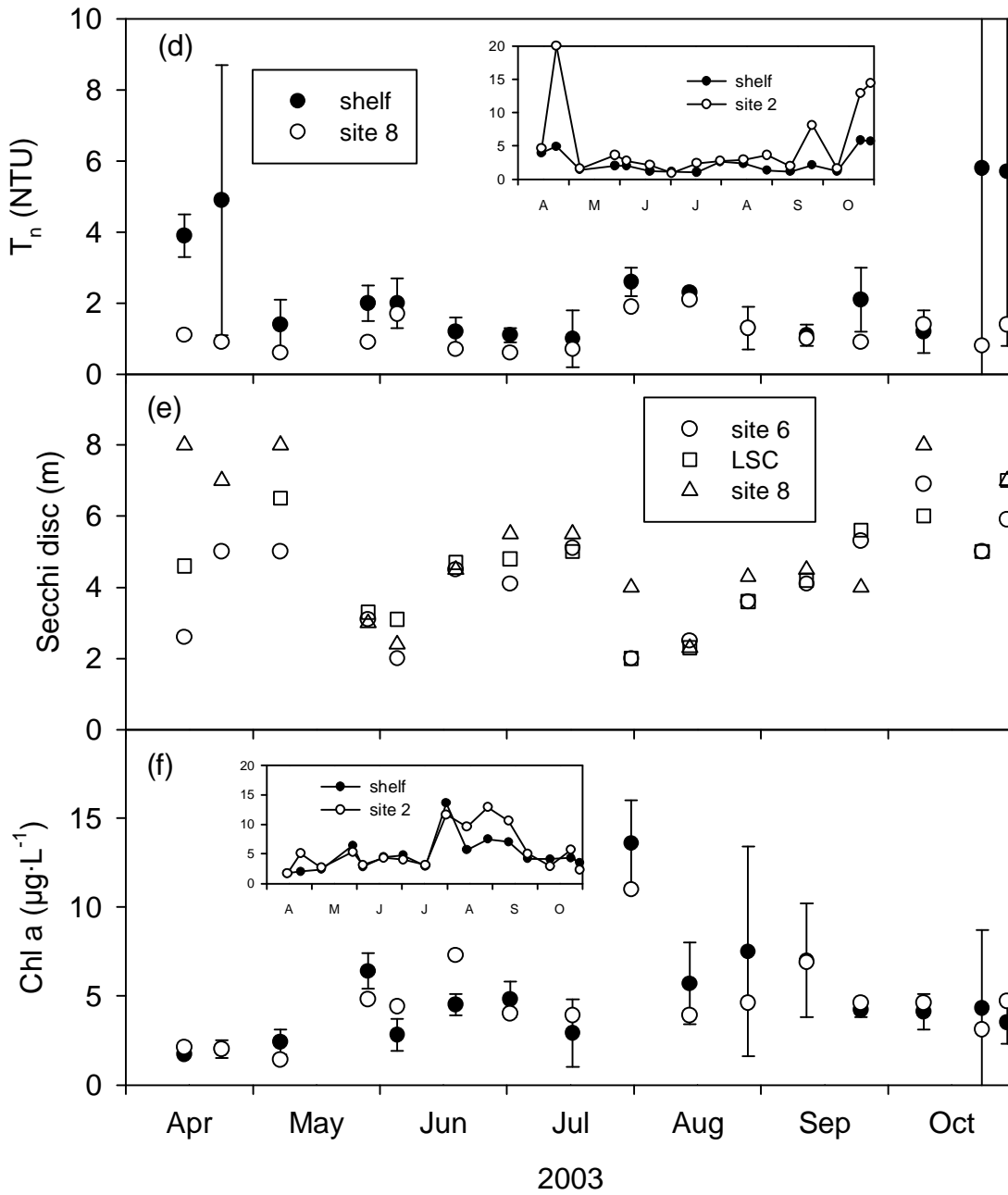


Figure 2d-f. Time series of parameter values for Cayuga Lake for 2002: (d) T_n , (e) Secchi disc, and (f) Chl a. Insets present results for site 2. Results for the “shelf” are averages; the error bars represent spatial variation with dimensions of ± 1 standard deviation.

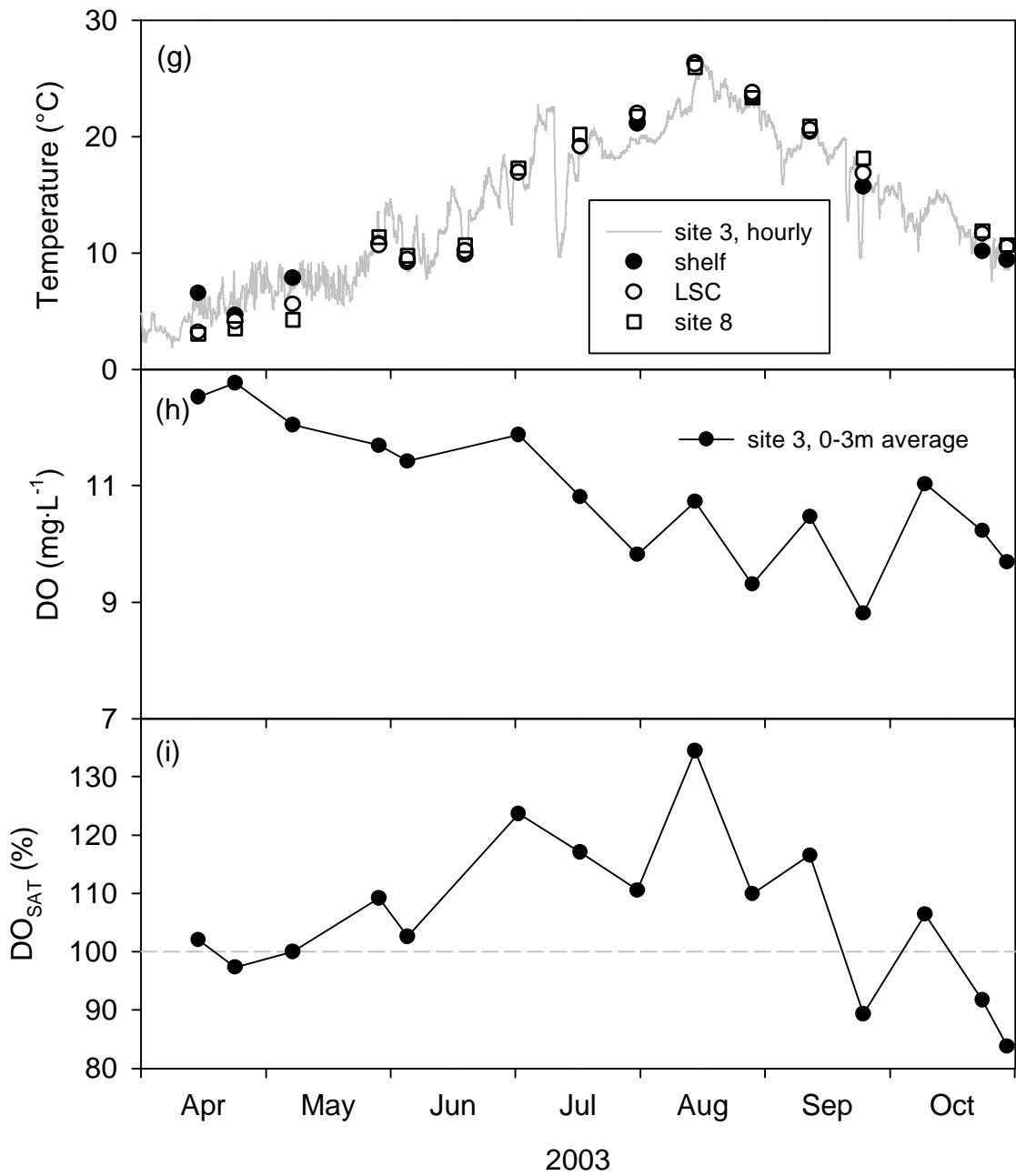


Figure 2g-i. Time series of parameter values for Cayuga Lake for 2003: (g) temperature, (h) DO, and (i) DO_{SAT}.

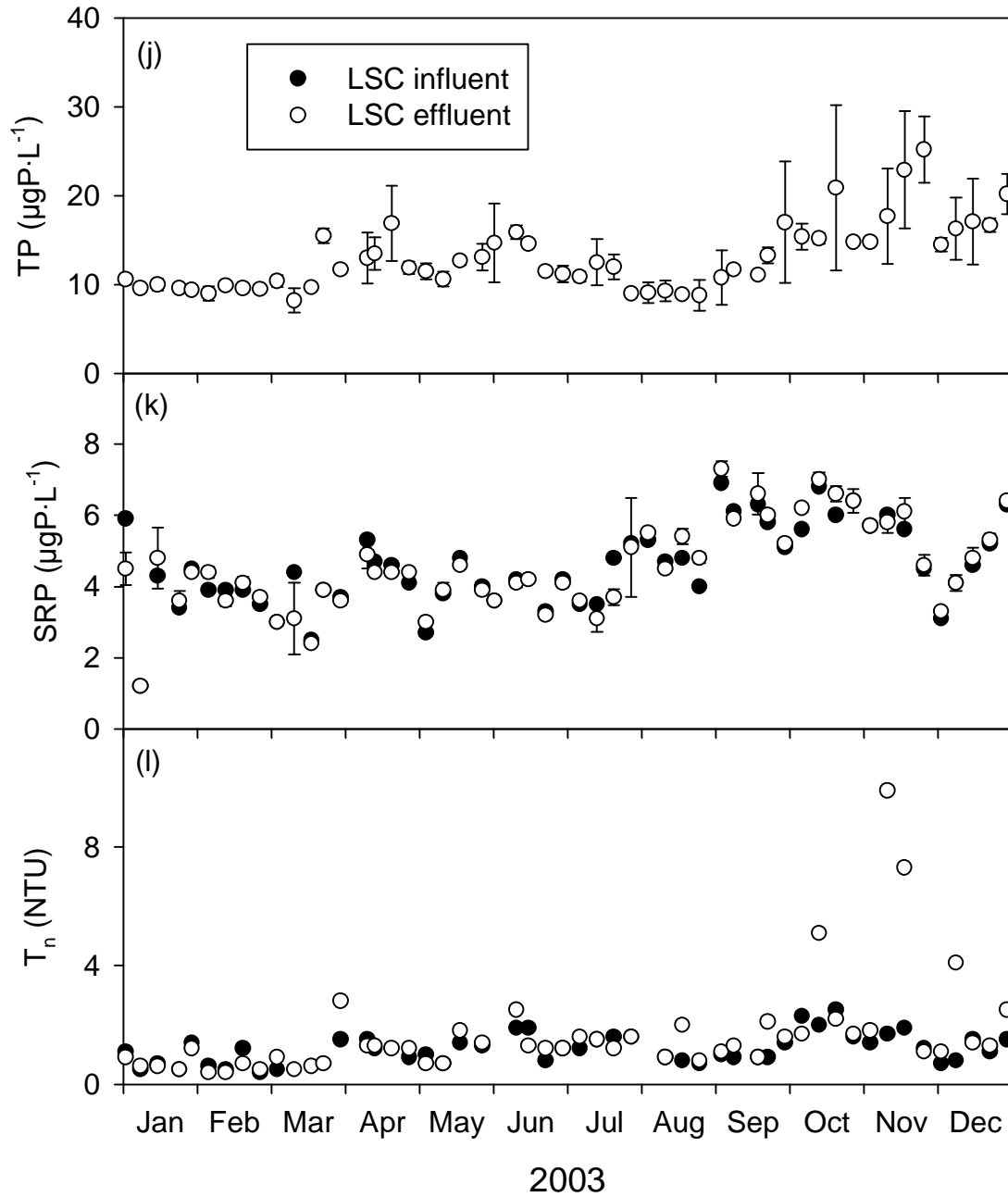


Figure 2j-1. Time series of parameter values for the LSC influent and effluent for 2003: (j) TP (not measured for LSC influent), (k) SRP, and (l) T_n . Error bars represent 95% confidence intervals determined from analyses of field triplicates.

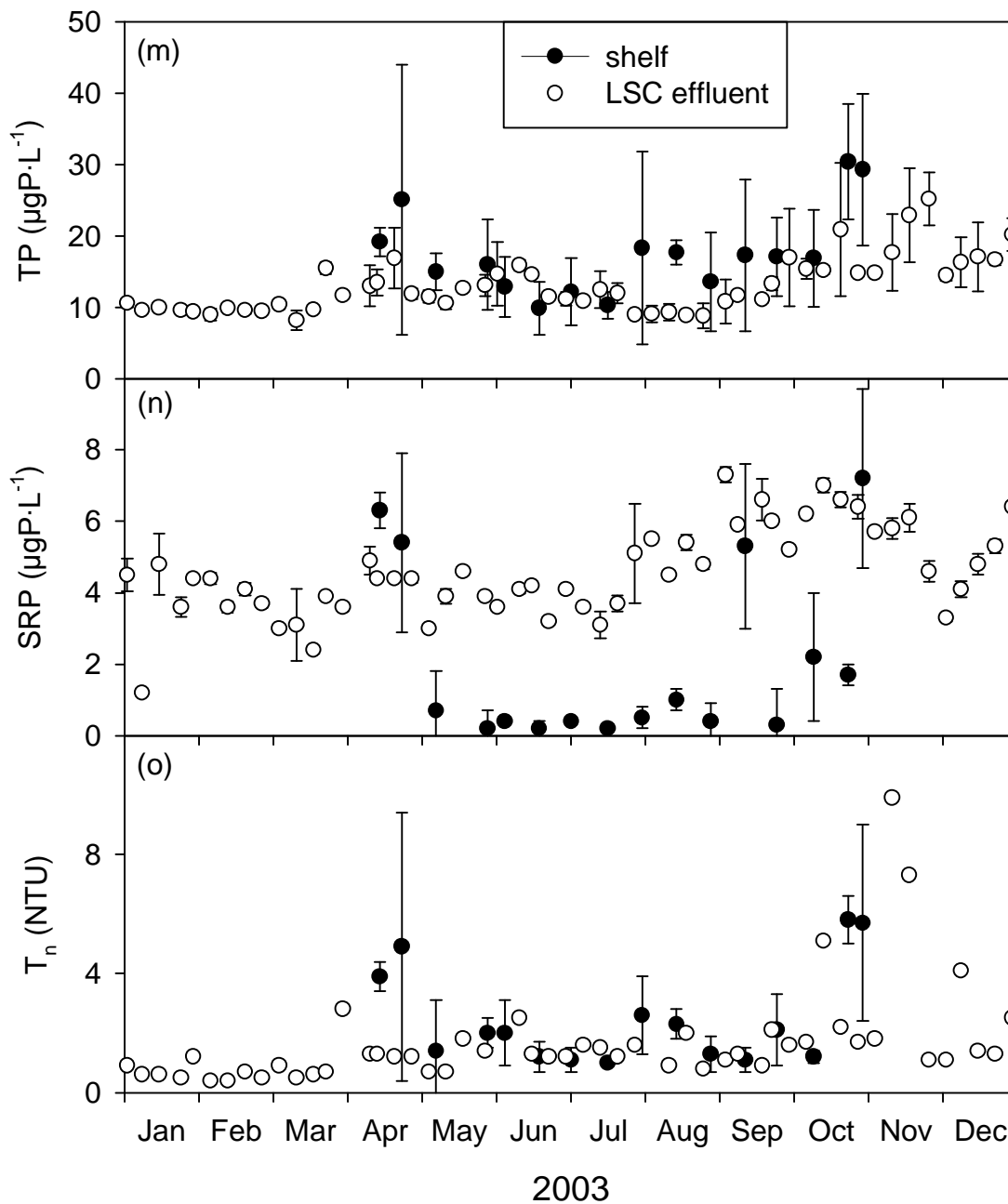


Figure 2m-o. Time series of parameter values for the south shelf and the LSC effluent for 2003: (m) TP, (n) SRP, and (o) T_n . Results for the “shelf” are averages; the error bars represent spatial variation with dimensions of ± 1 standard deviation. Error bars for the LSC effluent represent 95% confidence intervals determined from analyses of field triplicates.

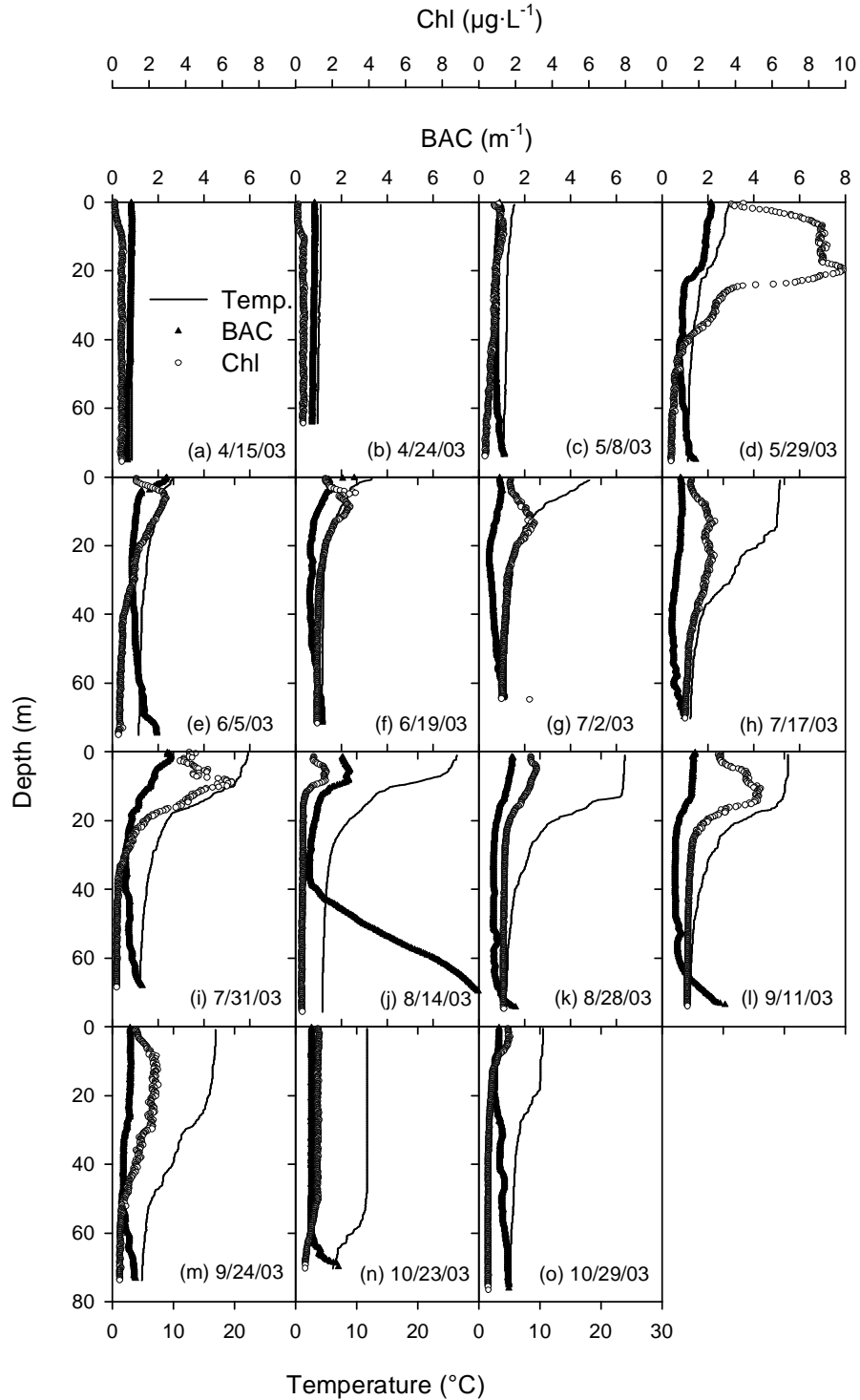


Figure 3. Vertical profiles of temperature, chlorophyll fluorescence, and beam attenuation coefficient (BAC) for LSC site in 2003: (a) April 15, (b) April 24, (c) May 8, (d) May 29, (e) June 5, (f) June 19, (g) July 2, (h) July 17, (i) July 31, (j) August 14, (k) August 28, (l) September 11, (m) September 24, (n) October 23, (o) October 29.

4. Selected Topics

4.1. Measures of Clarity

Secchi disc is a systematically flawed measure of clarity for much of the southern portion of Cayuga Lake monitored in this program because of its shallowness. Secchi disc transparency (SD) was observed to extend beyond the lake depth at sites 1, 2, 3, 4, 5 and 7 on several occasions during the 2003 study interval (see Appendix 1). Use of the population of SD measurements available (i.e., observations of $SD < \text{lake depth}$) results in systematic under-representation of clarity for each of these sites by eliminating the inclusion of deeper measurements. In addition, the SD measure is compromised as it approaches the bottom because reflection by the bottom rather than particles in the water can influence the measure. It may be prudent to consider an alternate representation of clarity that does not have these limitations. Turbidity (T_n) represents a reasonable alternative, in systems where particles regulate clarity (Effler 1988).

The relationship between SD and T_n is evaluated in the inverse format (e.g., Effler 1988) in Figure 4. A linear relationship is expected (Effler 1988), and has been observed for the observations made during this study (1998 – 2003; Figure 4). Based on these results (Figure 4), T_n should be considered as an alternate, and apparently more robust, measure of light penetration in shallow portions of the monitored area. The relationship between SD and T_n has remained consistent throughout the six study years.

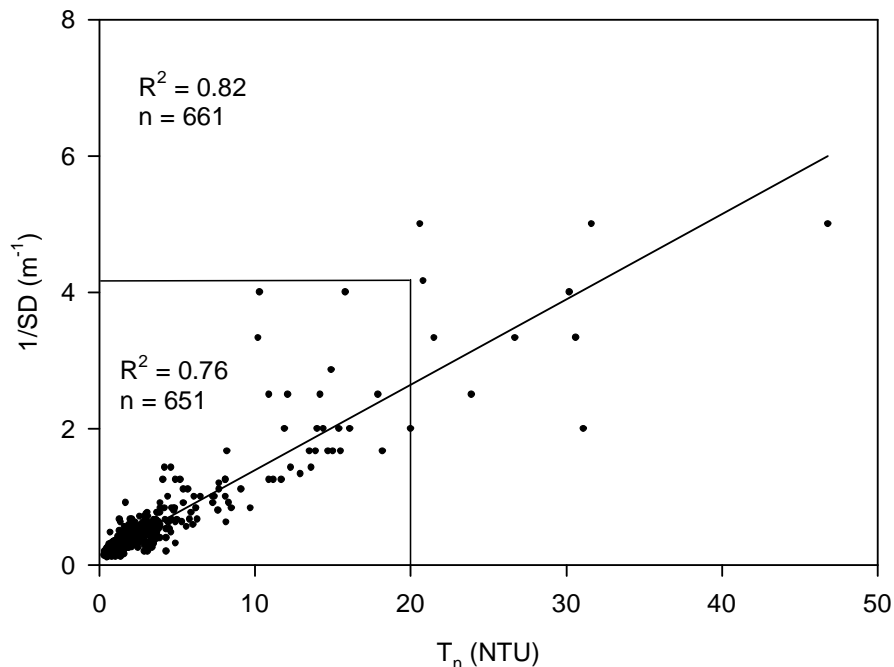


Figure 4. Relationship between Secchi disc transparency (SD) and turbidity in the southern end of Cayuga Lake based on paired measurements made during the 1998 – 2003 study interval.

4.2. Inputs of Phosphorus to Southern End of Cayuga Lake

Phosphorus loading is an important driver of primary production in phosphorus limited lakes. It is therefore valuable to consider the relative magnitudes of the various sources of phosphorus that enter the southern end of Cayuga Lake. Monthly average loading estimates are presented for the Ithaca Area and Cayuga Heights wastewater treatment plants (WWTPs) for the 1998 - 2003 interval (Table 7, Figure 5), based on flow and concentration data made available by these facilities. Discharge flows are measured continuously at these facilities. Concentrations of total phosphorus (TP) in the effluents are measured twice per week at the Ithaca Area WWTP and once per week at the Cayuga Heights WWTP. The estimates of the monthly loads (Table 7, Figure 5) are the product of the monthly average flows and concentrations. Other estimation techniques may result in modest differences in these loads. Rather wide monthly and interannual differences in loading rates have been observed for both WWTPs (Table 7, Figure 5) over the 1998 – 2003 interval. Phosphorus loading from the Ithaca Area WWTP was lower in 2003 than in the preceding five years (Table 7, Figure 5). The TP permit requirement is 40 pounds per day (18.1 kg per day) for the Ithaca Area WWTP and 1 mg·L⁻¹ for the Cayuga Heights WWTP.

Estimates of monthly tributary phosphorus loading presented in the **Draft Environmental Impact Statement** for the LSC facility, for the combined inputs of Fall Creek and Cayuga Inlet, for the May – October interval are included for reference in Table 7 and Figure 5. These were developed for what was described in that document as an “average hydrologic year”. The estimates were based on historic data for these two tributaries. Tributary loads can vary substantially year-to-year, based on natural variations in runoff. Further, the tributary phosphorus loads of Table 7 and Figure 5 were not for TP, but rather total soluble phosphorus [see Bouldin (1975) for analytical protocols], to better represent the potential for these inputs to support plant growth.

Estimates of monthly TP loading to the shelf from the LSC facility and the percent contribution of this source are presented in Table 7 and Figure 5. Concentrations of TP are measured weekly at the LSC discharge. The estimates of the monthly loads (Table 7, Figure 5) are the product of the monthly average flows and concentrations that are reported monthly as part of the Discharge Monitoring Report (DMR; Appendix 2). The average TP loading rate from LSC during the May – October period was 1.1 kg·d⁻¹, or 4.0% of the total TP load to the shelf. This is a smaller contribution than the 2.9 kg·d⁻¹, or 4.8% of the total TP load to the shelf, projected in the **Draft Environmental Impact Statement** for the LSC facility (Stearns and Wheler 1997).

Average levels of TP, SRP and T_n in the LSC effluent and on the shelf are presented in Table 8. Compared to the receiving waters of the shelf, TP concentrations in the LSC effluent were on average about 5 µg·L⁻¹ lower and SRP concentrations were about 3 µg·L⁻¹ higher (Table 8). Average turbidity levels were 1.0 NTU higher on the shelf than in the LSC effluent. Levels of TP, SRP and T_n varied widely over time and space on the shelf during 2003 (Figure 2p-r). In contrast, the LSC effluent phosphorus concentrations and turbidity levels remained relatively uniform. Levels of TP and T_n in the LSC effluent were elevated on several sampling days during the October – December interval of 2003 (Figure 2j, l). The cause(s) of these higher values is unknown.

Table 7: Estimates of monthly external loads of phosphorus to the southern portion of Cayuga Lake, 1998 - 2003.

Month	Ithaca Area WWTP ^a (kg·d ⁻¹)						Cayuga Heights WWTP ^b (kg·d ⁻¹)					
	1998	1999	2000	2001	2002	2003	1998	1999	2000	2001	2002	2003
May	14.1	19.7	24.1	15.8	12.4	11.0	8.7	3.7	3.5	5.5	4.4	2.7
June	5.8	9.1	16.6	11.2	7.9	6.0	7.5	4.3	5.1	4.0	3.5	7.8
July	16.4	11.4	13.7	15.2	10.4	8.5	4.4	2.6	3.4	4.2	3.8	3.9
August	17.0	12.5	19.1	15.2	16.2	13.8	4.7	1.5	4.6	7.1	2.0	3.1
September	32.8	20.0	18.5	22.0	11.4	11.9	7.7	1.8	4.0	6.6	2.8	3.4
October	16.2	9.4	15.4	16.4	13.6	14.5	9.1	1.7	4.1	2.8	3.1	5.3
<i>Mean</i>	<i>17.1</i>	<i>13.7</i>	<i>16.5</i>	<i>16.0</i>	<i>12.0</i>	<i>11.0</i>	<i>7.0</i>	<i>2.6</i>	<i>4.1</i>	<i>5.0</i>	<i>3.3</i>	<i>4.4</i>

Month	Tributary ^c (kg·d ⁻¹) average year	LSC ^d (kg·d ⁻¹)				Total (kg·d ⁻¹)				% LSC			
		2000	2001	2002	2003	2000	2001	2002	2003	2000	2001	2002	2003
May	29.0	-	0.7	0.6	0.6	56.6	51.0	46.4	43.3	-	1.4	1.3	1.4
June	15.8	-	1.1	1.0	1.2	37.5	32.1	28.2	30.8	-	3.4	3.5	3.9
July	8.8	1.4	1.0	1.8	1.2	27.3	29.2	24.8	22.4	5.1	3.4	7.3	5.4
August	6.0	1.0	1.4	1.2	1.2	30.7	29.7	25.4	24.1	3.3	4.7	4.7	5.0
September	7.5	0.9	1.0	1.0	1.3	30.9	37.1	22.7	24.1	2.9	2.7	4.4	5.4
October	13.1	0.6	0.7	0.7	0.9	33.2	33.0	30.5	33.8	1.8	2.1	2.3	2.7
<i>Mean</i>	<i>13.3</i>	<i>1.0</i>	<i>1.0</i>	<i>1.1</i>	<i>1.1</i>	<i>34.9</i>	<i>35.4</i>	<i>29.7</i>	<i>29.8</i>	<i>2.9</i>	<i>3.0</i>	<i>3.5</i>	<i>4.0</i>

^a total phosphorus; personal communication with Jose Lozano, Director

^b total phosphorus; from USEPA website-http://www.epa.gov/enviro/index_java.html

^c total soluble phosphorus, for average hydrologic year; summation of Fall Creek and Cayuga Inlet; from Draft Environmental Impact Statement, LSC Cornell University, 1997

^d total phosphorus; from facility permit reporting

Table 8: Average values and standard deviations for TP, SRP, and T_n in the LSC effluent and on the shelf. Averages determined from observations made during the April – October interval of 2003.

Location	TP (µg·L ⁻¹)	SRP (µg·L ⁻¹)	T _n (NTU)
LSC effluent (n = 30)	12.7±2.8	4.9±1.2	1.5±0.8
Shelf (n = 16)	17.6±6.1	2.0±2.5	2.5±1.7

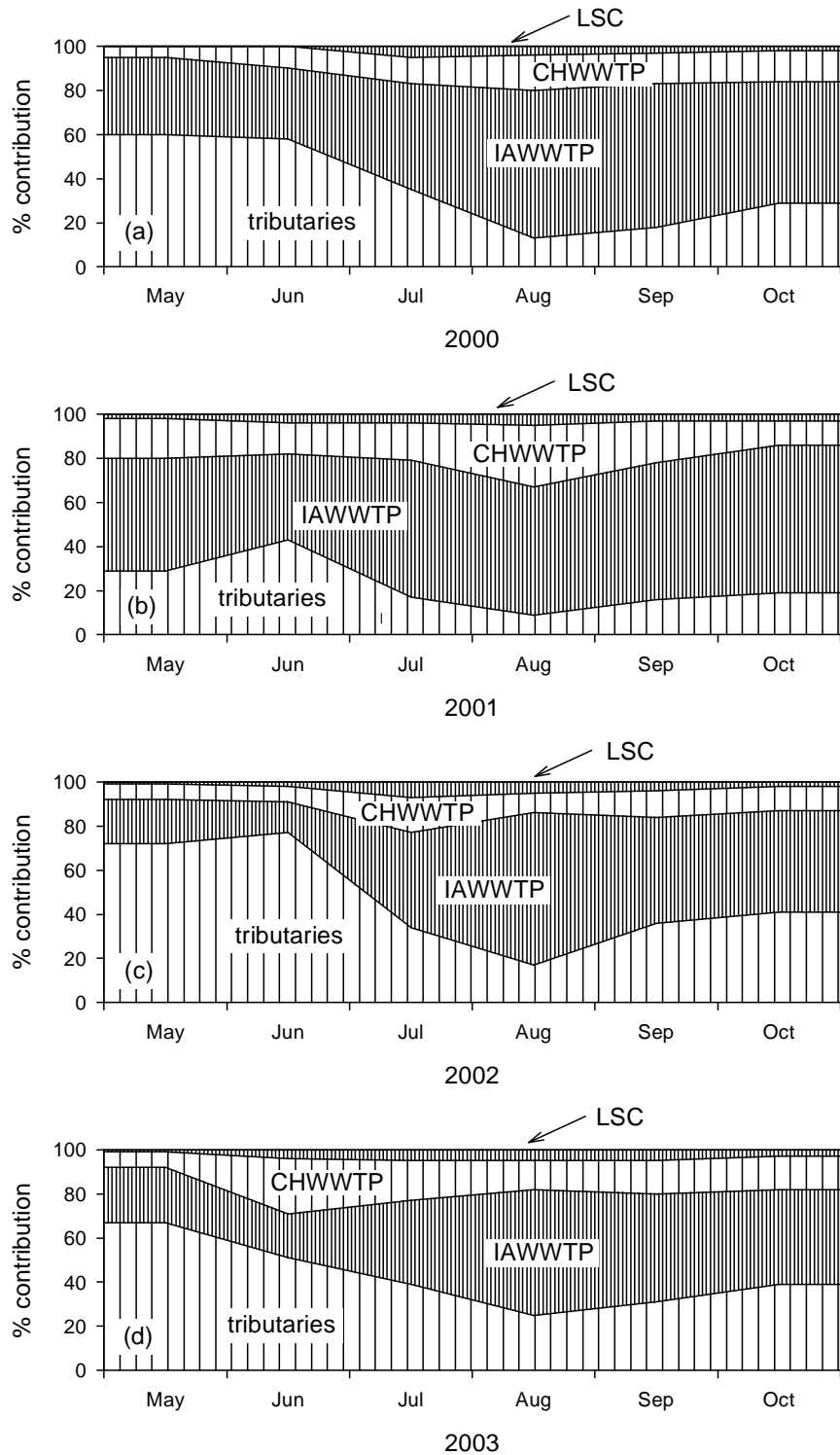


Figure 5. Time series of estimated monthly external loads of phosphorus to the southern portion of Cayuga Lake, partitioned according to source: (a) 2000, (b) 2001, (c) 2002, and (d) 2003. Loads are for total phosphorus with the exception of tributary loading, which is for total soluble phosphorus.

4.3. Variations in Runoff and Wind Speed

Meteorological conditions and coupled features of runoff have important effects on lake ecosystems. These conditions are not subject to management, but in fact demonstrate wide variations in many climates that can strongly modify measures of water quality (e.g., Auer and Effler 1989, Lam et al. 1987). Thus the effects of natural variations in these conditions can be mistaken for impacts of man's activities (e.g., pollution). The setting of the southern end of the lake, including the localized entry of tributary flows and its shallowness, may promote interpretive interferences with the measurements of total phosphorus (TP), Secchi disc transparency (SD), and turbidity (T_n). These interferences are associated with potential influxes of non-phytoplankton particles that would diminish SD and increase T_n and TP concentrations, features that could be misinterpreted as reflecting increases in phytoplankton concentrations. These influxes may be associated with external loads carried by the tributaries, particularly during runoff events, and internal loads associated with sediment resuspension, driven by wind events (e.g., Bloesch 1995). Thus, it is prudent to consider natural variations in tributary flow and wind speed in evaluating seasonal and interannual differences in these parameters for the southern end of Cayuga Lake. Interannual variations in runoff and wind speed are discussed in **Section 4.7 – Interannual Comparisons**, and illustrated in Figures 11 and 13.

Runoff and wind conditions for the study period of 2003 are represented here by daily average flows measured in Fall Creek by USGS, and daily average wind speed, out of the north to northwest, measured by Cornell University (Figure 6). These conditions are placed in a historic perspective by comparison to available records. Fall Creek has been reported to be a good indicator of lake-wide runoff conditions (Effler et al. 1989). The record for Fall Creek is quite long, nearly 80 years; the wind database reflects 21 years of measurements. Daily average measurements of Fall Creek flow and wind speed for 2003 are compared to time-series of daily median values for the available records (Figure 6a and c). Additionally, monthly average flows for the study period are compared to quartiles for the period of record (Figure 6b). Due to the orientation of the southern end of Cayuga Lake, winds out of the north to northwest ($315^\circ - 360^\circ$) are expected to drive the greatest turbulence, and thus resuspension, in this part of the lake. However, if seiche action is a major cause of sediment resuspension a south wind will also be important.

Major runoff events (Fall Creek flow >800 cfs) occurred in early April, early June, late July, and late October of 2003 (Figure 6a). In addition, a number of smaller storms were scattered throughout the April-October interval (Figure 6a). In-lake sampling was conducted during or immediately following several of these runoff events (Figure 6a). For example, the October 29 sampling occurred during a major runoff event (Figure 6a). When compared to the historic record, Fall Creek flows were elevated for much of the April – October interval of 2003 (Figure 6b). Monthly average flows were near the long-term average in April and below the 25-percentile level in May (Figure 6b). Fall Creek flows were particularly high from June through October, exceeding the 75-percentile level in each of these months (Figure 6b).

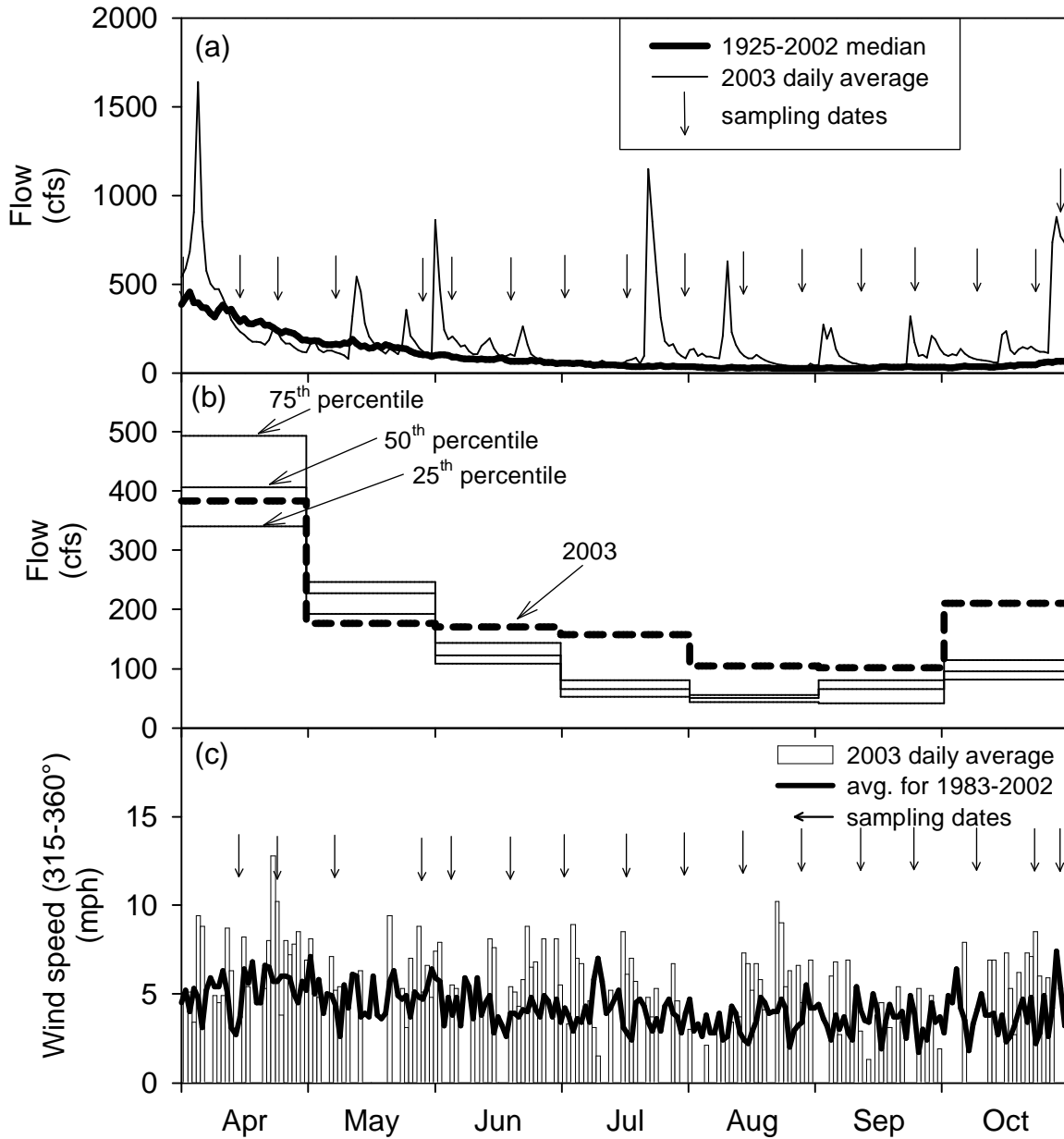


Figure 6. Runoff and wind conditions for the April – October interval of 2003: (a) daily average flows in Fall Creek, compared to median daily values for the 1925 – 2002 record, (b) monthly flows in 2003 compared, to quartile levels of flow for the 1925 – 2002 record, and (c) daily average wind speed out of the north to northwest, compared to median values for the 1983 – 2002 record.

Major wind events (e.g., protracted intervals of high winds) from the north to northwest did not occur over the study interval of 2003 (Figure 6c). However, winds were above average for extended periods during late April, mid to late August, and mid-October (Figure 6c). Wind velocities were distinctly above average on, or before, the monitoring days of April 24, July 17, and October 23 (Figure 6c).

4.4. Limitations in Measures of Trophic State on the Shelf

Recurring circumstantial scientific evidence, provided by the findings of five consecutive study years (Upstate Freshwater Institute 1999, 2000, 2001, 2002, 2003) has demonstrated that T_n and TP are systematically flawed indicators of the trophic state on the shelf. In particular, substantial variations and increases in both parameters on the south shelf appear to be uncoupled at times from patterns and magnitudes of phytoplankton biomass. These features appear to be associated with greater contributions of non-phytoplankton particles (e.g. clay and silt) to the measures of TP and T_n on the south shelf. Four lines of circumstantial evidence supporting this position have been presented, based on observations from the 1998 - 2002 study years (Upstate Freshwater Institute 1999, 2000, 2001, 2002, 2003). Observations from the 2003 study year provide additional evidence that T_n and TP are compromised as trophic indicators in this system because of the contributions of inanimate non-phytoplankton particles (tripton):

1. the highest T_n (Figure 2d) values observed during 2003 for the shelf were observed following major runoff events in April and October (Figure 6a). This suggests greater contributions of non-phytoplankton particles to the measurements of T_n following runoff events.
2. elevated T_n values were reported for the 1999, 2000 and 2002 study years (Upstate Freshwater Institute 2000, 2001, 2003) at the deep water sites during “whiting” events in late July and August. These increases in T_n were driven largely by increases in T_c (calcium carbonate turbidity; Figure 7). The increase in T_c observed during August 2003 is suggestive of a “whiting” event (Figure 7).
3. the ratio of particulate P (PP) to chlorophyll **a** was often substantially higher on the south shelf than at the deep stations (Figure 8), suggesting greater contributions of non-phytoplankton particles to the PP pool at the southern end of the lake. Further, unlike the deep sites, the ratio was often above the range of values commonly associated with phytoplankton biomass (e.g., Bowie et al. 1985).
4. application of reasonable literature values of light scattering (e.g., T_n) per unit chlorophyll (e.g., Weidemann and Bannister 1986) to the chlorophyll **a** observations indicate that non-phytoplankton particles made greater contributions to T_n on the shelf than in deep waters (Figure 9). Non-phytoplankton particles were responsible for the high T_n levels on the shelf following major runoff events in April and October (Figure 9).

The 2003 results demonstrate substantial seasonal variations continue to occur for TP and T_n on the shelf that are uncoupled from the trophic state issue. Additional measurements were made in 1999 and 2000, beyond the scope of the LSC monitoring program, to more comprehensively resolve the constituents/processes regulating the SD and TP measurements (Effler et al. 2002). Effler et al. (2002) demonstrated that inorganic particles (primarily clay minerals, quartz and calcium carbonate), rather than phytoplankton, are the primary regulators of clarity, represent most of the PP, and are responsible for the higher T_n , lower SD, and higher TP on the shelf compared to deeper portions of the lake.

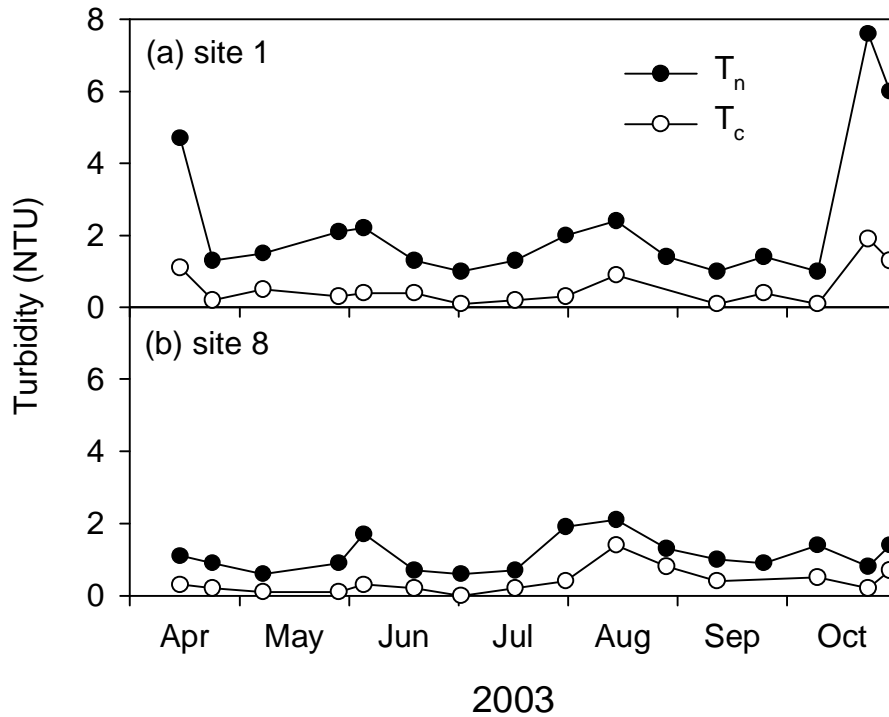


Figure 7. Distributions of total turbidity (T_n) and calcium carbonate turbidity (T_c) in the upper waters of Cayuga Lake in 2003: (a) site 1, (b) site 8.

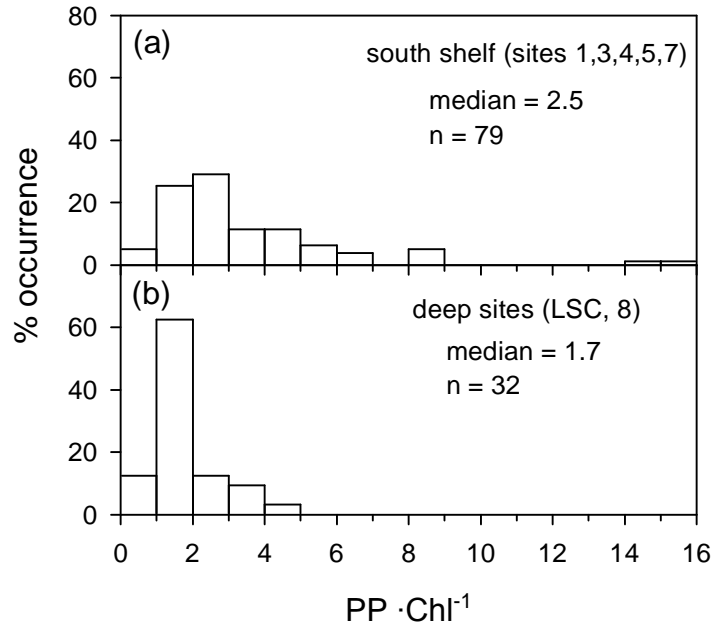


Figure 8. Distributions of the particulate P (PP) to chlorophyll a ratio values in Cayuga Lake in 2003: (a) south shelf sites, and (b) deep water sites.

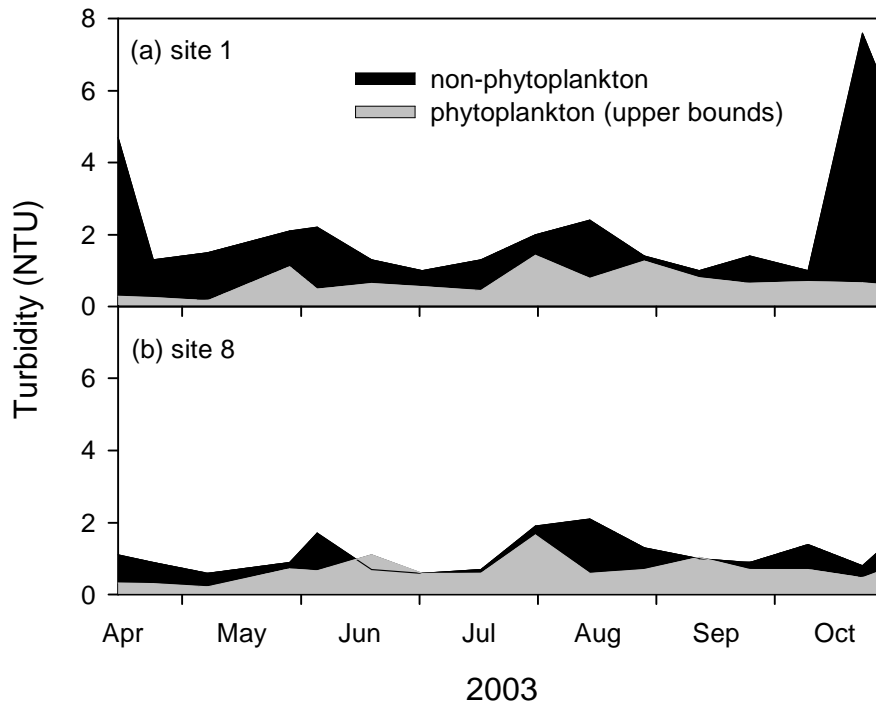


Figure 9. Time-series of T_n and contributions from components (phytoplankton and non-phytoplankton) for the April – October interval of 2003: (a) site 1, and (b) site 8.

4.5. Continuation of the Long-Term Record of Water Quality/Eutrophication Indicators

Systematic changes in water quality can only be quantitatively documented if reliable measurements are available for historic conditions. Concentrations of TP and chlorophyll **a** have been measured irregularly in the open waters of Cayuga Lake over the last three decades. Measurements made over the late 1960s to mid 1970s were made mostly as part of research conducted by Cornell University staff (Tables 9 and 10). These data were collected mostly at deep water locations. No comprehensive data sets were found to represent conditions in the 1980s. Measurements were continued in the 1994 – 1996 interval as part of studies conducted to support preparation of the **Draft Environmental Impact Statement** for the LSC facility (Stearns and Wheler 1997). These included observations for both the shelf and deeper locations (Tables 9 and 10). The record continues to be updated annually, for both a deep water location and the shelf, based on monitoring sponsored by Cornell University related to operation of the LSC facility (1998 – 2003, documented here).

Summer (June – August) average concentrations are presented for the lake's upper waters; sources of data are included (Tables 9 and 10). Higher TP concentrations were observed on the shelf compared to deeper portions of the lake in all years monitored since 1994 (Table 9). Distinctly higher chlorophyll **a** concentrations were observed on the shelf in the summers of 1994 – 1996 compared to deeper water sites, however, the averages were similar over the 1998 – 2002 interval (Table 10). The 1998 average does not include June observations. Summer average concentrations of TP and chlorophyll **a** for deep water sites are consistent with a mesotrophic trophic state classification (i.e., intermediate level of primary productivity; e.g., Chapra and Dobson 1981, Dobson et al. 1974, Vollenweider 1975).

Table 9: Summer (June - August) average total phosphorus (TP) concentrations for the upper waters of Cayuga Lake. June – September averages are included in parentheses for the 1998 – 2003 study years.

Year	Total Phosphorus ($\mu\text{g}\cdot\text{L}^{-1}$)		Source
	Deep-Water Location(s)	Southern Shelf	
1968 ^Δ	20.2 (n = 19)	-	Peterson 1971
1969 ^Δ	15.3 (n = 22)	-	Peterson 1971
1970 ^Δ	14.0 (n = 32)	-	Peterson 1971
1972 ^x	18.8 (n = 22)	-	USEPA 1974
1973 ^Δ	14.5 (n = 88)	-	Godfrey 1973
1994 ^{*⊕}	21.7	30.8	Stearns and Wheler 1997
1995 ^{*⊗}	16.5	23.7	Stearns and Wheler 1997
1996 ^{*⊗}	12.4	21.7	Stearns and Wheler 1997
1998 ⁺	14.7 (14.7)	26.5 (24.7)	UFI 1999
1999 ⁺⁺	10.6 (9.8)	15.9 (14.5)	UFI 2000
2000 ⁺⁺	11.9 (11.6)	19.4 (18.7)	UFI 2001
2001 ⁺⁺	14.0 (14.2)	21.4 (20.4)	UFI 2002
2002 ⁺⁺	14.7 (14.1)	22.1 (22.2)	UFI 2003
2003 ⁺⁺	10.2 (10.4)	13.6 (14.4)	this report

^Δ Myers Point

^x one sample, multiple sites and depths

^{*} averages of 0 m observations

⁺ July – August, 0 – 4 m composite samples

⁺⁺ 0 – 4 m composite samples

[⊕] site in 62 m of water, south of Myers Point, surface samples

[⊗] site in 70 m of water, south of Myers Point, surface samples

Table 10: Summer (June – August) average chlorophyll **a** concentrations for the upper waters of Cayuga Lake. June – September averages are included in parentheses for the 1998 – 2003 study years.

Year	Chlorophyll a ($\mu\text{g}\cdot\text{L}^{-1}$)		Source
	Deep-Water Location(s)	Southern Shelf	
1966*	2.8	-	Hamilton 1969
1968**	4.3	-	Wright 1969
1968 – 1970	4.8	-	Oglesby 1978
1970	3.7	-	Trautmann et al. 1982
1972	10.3	-	Oglesby 1978
1973	8.2	-	Trautmann et al. 1982
1974	8.1	-	Trautmann et al. 1982
1977	8.6	-	Trautmann et al. 1982
1978	6.5	-	Trautmann et al. 1982
1994	5.5	8.9	Stearns and Wheler 1997
1995	4.8	6.8	Stearns and Wheler 1997
1996	3.4	7.6	Stearns and Wheler 1997
1998 ⁺	4.8 (4.8)	5.7 (5.2)	UFI 1999
1999 ⁺⁺	4.7 (4.6)	4.4 (4.2)	UFI 2000
2000 ⁺⁺	4.8 (4.7)	5.5 (5.4)	UFI 2001
2001 ⁺⁺	4.7 (4.5)	4.6 (4.4)	UFI 2002
2002 ⁺⁺	5.1 (5.2)	4.8 (5.6)	UFI 2003
2003 ⁺⁺	5.6 (5.6)	6.0 (5.9)	this report

* Hamilton 1969, 15 dates

** Wright 1969, 4 dates – 7 to 9 longitudinal sites

⁺ July – August, 0 – 4 m composite samples

⁺⁺ 0 – 4 m composite samples

4.6. Comparison to Other Finger Lakes: Chlorophyll **a**

Synoptic surveys of all eleven Finger Lakes have been conducted in recent years (NYSDEC, with collaboration of the Upstate Freshwater Institute) that support comparison of selected conditions among these lakes. Chlorophyll **a** data (Callinan 2001) collected from those surveys are reviewed here, as this may be the most representative indicator of trophic state of the measurements made. Samples (n=15 to 16) were collected in these surveys over the spring to early fall interval of 1996 through 1999. The sample site for Cayuga Lake for this program coincides approximately with site 8 of the LSC monitoring program (Figure 1b).

There is not universal agreement on the concentrations of chlorophyll **a** that demarcate trophic states. A summer average value of $2.0 \mu\text{g}\cdot\text{L}^{-1}$ has been used as the demarcation between oligotrophy and mesotrophy (Dobson et al. 1974, National Academy of Science 1972). There is less agreement for the demarcation between mesotrophy and eutrophy; the boundary summer average value reported from different sources (e.g., Dobson et al. 1974, National Academy of Science 1972, Great Lakes Group 1976) ranges from 8 to $12 \mu\text{g}\cdot\text{L}^{-1}$.

The average chlorophyll *a* concentration for Cayuga Lake for this synoptic program ($3.5 \mu\text{g}\cdot\text{L}^{-1}$) is compared to the values measured in the other ten Finger Lakes in Figure 10. These data support Cayuga Lake's classification as mesotrophic. Six of the lakes had average concentrations lower than observed for Cayuga Lake (Figure 10). Two of the lakes, Canandaigua and Skaneateles, had concentrations consistent with oligotrophy, while two (Conesus and Honeoye) bordered on eutrophy (Figure 10).

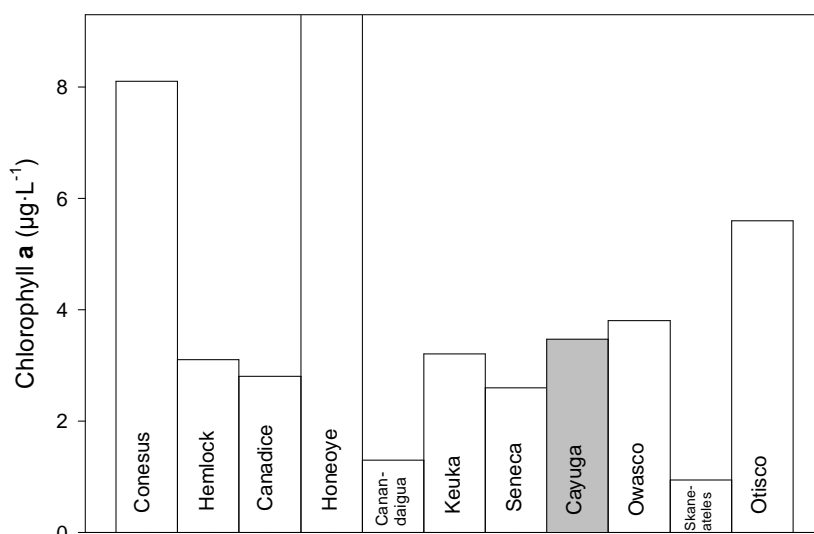


Figure 10. Comparison of average chlorophyll *a* concentrations for the spring-early fall interval for the eleven Finger Lakes, based on samples ($n=15$ to 16) collected over the 1996 through 1999 interval (data from Callinan 2001).

4.7. Interannual Comparisons

Interannual differences in water quality can occur as a result of both human interventions and natural variations in climate. Because of its location and shallowness, water quality on the south shelf can vary substantially from year to year as a result of changes in forcing conditions. Conditions for runoff, wind speed and summed TP loading from the Ithaca Area WWTP, Cayuga Heights WWTP and the LSC facility, for 2003 are compared here to the five previous study years (1998 – 2002; Figure 11). When compared to flow conditions of the preceding five years, the most striking features of the Fall Creek hydrograph for 2003 are the unusually large runoff events in July, August and October (Figure 11a). Due to the timing of the fixed-frequency sampling program (Figure 6a), water quality impacts were observed clearly for the October storm, but not for the July and August storms (e.g., Figure 2d).

Daily average wind speeds, out of the north to northwest, for the 1998 - 2002 interval and the 2003 study period are presented in Figure 11b for comparison. Major year-to-year differences have not been observed for this metric (Figure 11b). However, the two extended periods of light

winds observed in 2003 (one during late July and early August and the other during September; Figure 11b) represent unusual conditions for this location. Estimates of monthly average total phosphorus (TP) loads from point sources are compared here for 1998 - 2002 and 2003 (Figure 11c). Total phosphorus loading from point sources was relatively low in 2003 compared to the previous five study years. In fact, TP loads for May and July of 2003 were the lowest of the 1998 – 2003 record.

Time series of TP, Chl, and T_n are presented for the April – October interval of the six study years (Figure 12). Data were not collected during the April – June interval of 1998. Plotted values (the mean of observations for sites 3, 4, 5, and the average of sites 1 and 7) are intended to represent conditions on the shelf. Concentrations of TP were lower during 2003 than in the previous five study years (Figure 12a, Table 9). TP concentrations $> 20 \mu\text{g}\cdot\text{L}^{-1}$ were only observed during high runoff intervals in April and October (Figure 6a, Figure 12a). High TP concentrations (e.g., $> 30 \mu\text{g}\cdot\text{L}^{-1}$) were not observed during the study intervals of 1999 or 2003.

Chlorophyll **a** concentrations were similar during the six study years with the exception of the higher values observed during the July – August intervals of 2000 and 2003 and the spike in early September of 2002 (Figure 12b). In general, chlorophyll **a** concentrations have been lowest during spring and highest during mid-summer (Figure 12b). High turbidity values were observed on sampling dates that coincided with major runoff events in early July 1998, early April 2000, mid-June 2000, early April 2001, and late June 2001 (Figure 12c). The highest turbidity values measured in 2003 were associated with high runoff intervals during April and October (Figure 12c). High turbidity values (e.g., > 5 NTU) were not observed during the 1999 study interval.

The temporally detailed data presented in Figures 11 and 12 are summarized in Figure 13 as box plots for the six study years. The dimensions of the boxes are identified according to the key located to the right of Figure 13a. Fall Creek flows were highest in 2000; runoff was also relatively high in 2002 and 2003 (Figure 13a). Flows were lowest for the study intervals of 1999 and 2001. Average wind speeds were essentially equal for the 1998 – 2002 study intervals, and slightly lower in 2003 (Figure 13b). Total phosphorus loading from point sources was relatively low in 1999, 2002 and 2003, and higher in 1998, 2000 and 2001 (Figure 13c). Greater month-to-month variability in TP loading was observed in 1998 than in the other study years (Figure 13c). Study period medians for TP, Chl and T_n on the shelf were lowest for 1999, the driest of the study years (Figure 13d-f). Temporal variability for these three metrics was also much lower during the 1999 study interval (Figure 13d-f).

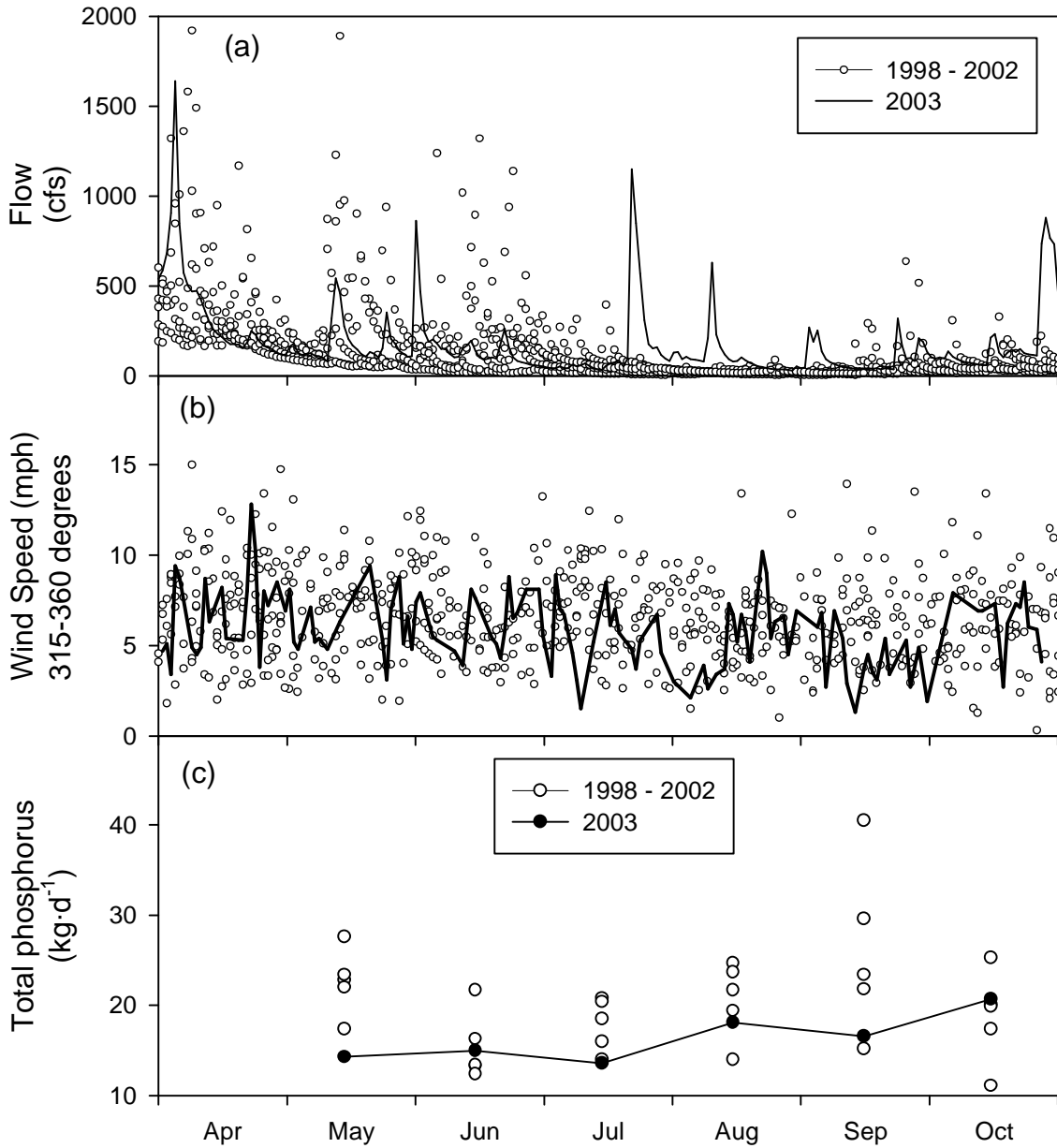


Figure 11. Comparison of 1998, 1999, 2000, 2001, 2002 and 2003 conditions for runoff, wind and total phosphorus loading for the April – October interval: (a) daily average flows in Fall Creek, (b) daily average wind speed, and (c) summed monthly loads of total phosphorus (TP) to southern Cayuga Lake from the Ithaca Area WWTP, Cayuga Heights WWTP, and the LSC facility.

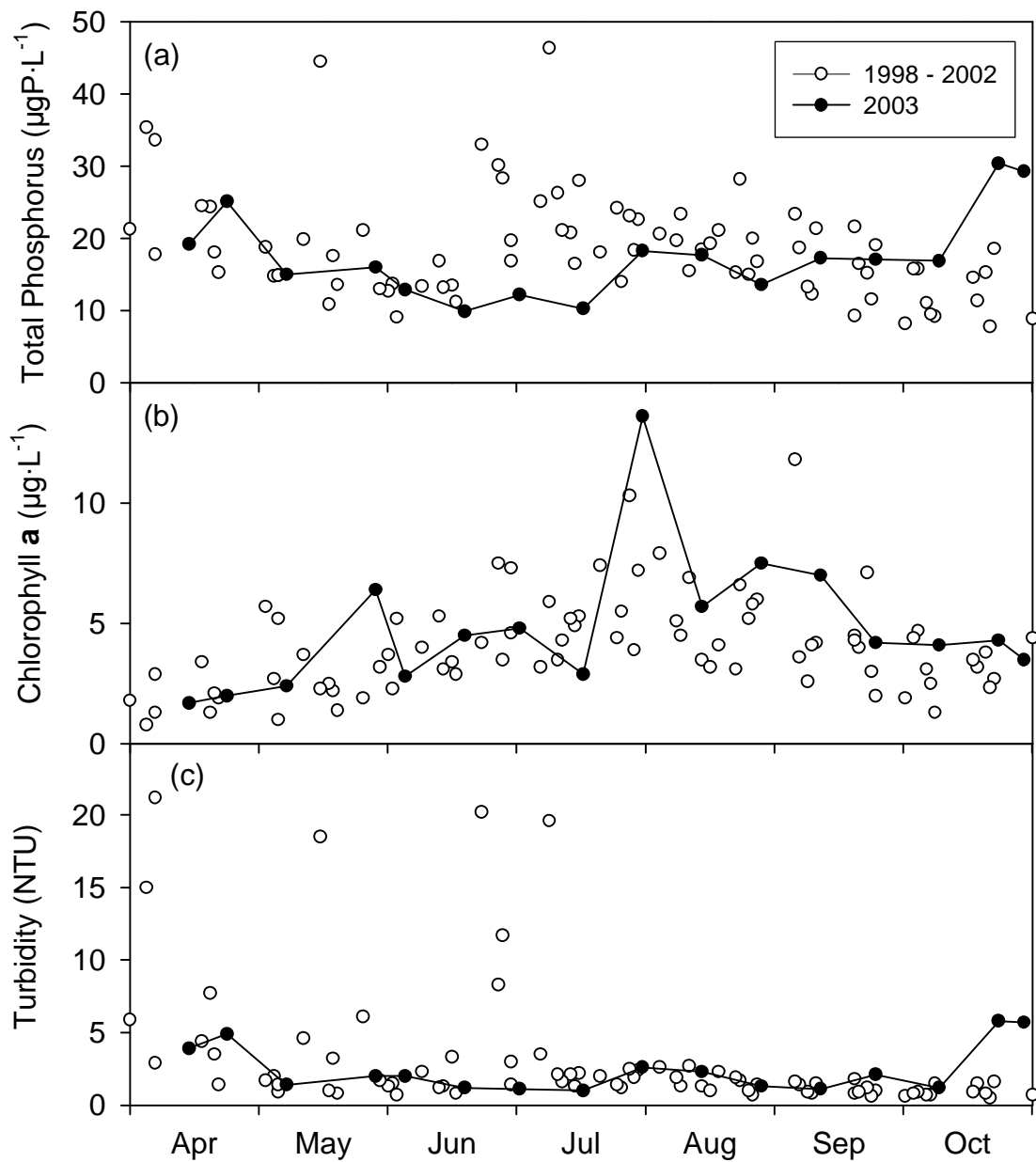


Figure 12. Comparison of 1998, 1999, 2000, 2001, 2002 and 2003 conditions for total phosphorus, chlorophyll **a**, and turbidity on the south shelf of Cayuga Lake for the April – October interval: (a) total phosphorus (TP), (b) chlorophyll **a**, and (c) turbidity (T_n).

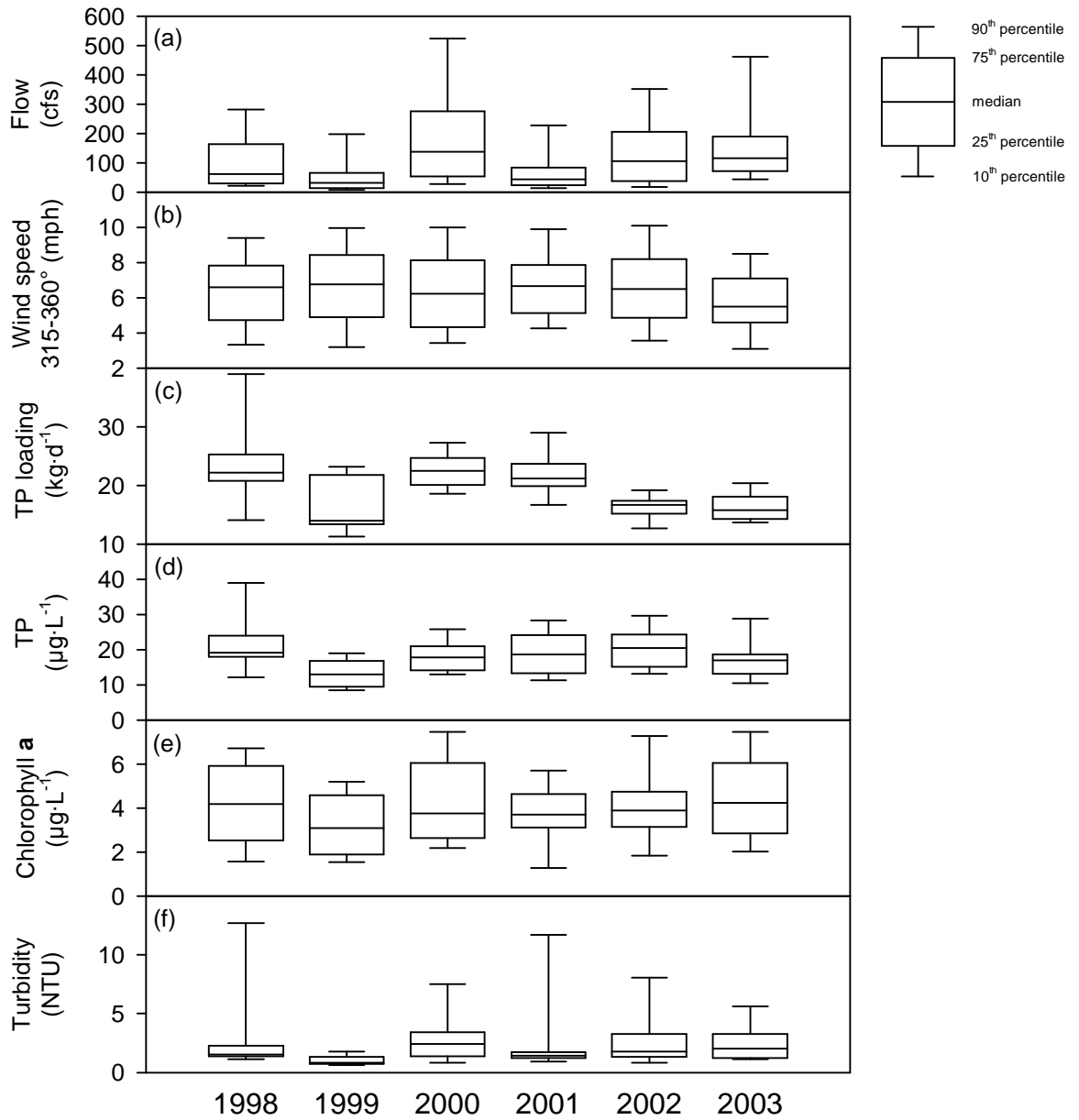


Figure 13. Comparison of study interval averages for runoff, wind, total phosphorus loading, total phosphorus concentration, chlorophyll **a** concentration and turbidity: (a) Fall Creek flow, (b) wind speed, (c) summed loads of total phosphorus (TP) from the Ithaca Area WWTP, Cayuga Heights WWTP and the LSC facility, (d) total phosphorus concentration on the south shelf, (e) chlorophyll **a** concentration on the south shelf, and (f) turbidity on the south shelf. 1998 averages for total phosphorus concentration, chlorophyll **a** concentration and turbidity are for the July – October interval; all other averages are for the April – October interval.

4.8. Noteworthy Observations from the 2003 Data

1. site 2 was enriched in all three forms of phosphorus (TP, TDP, and SRP) compared to the other monitored sites, particularly during the August – October interval (Figure 2, Table 6).
2. site 7 had the highest average levels of chlorophyll **a** (Chl) and turbidity (T_n) and was enriched in all three forms of phosphorus (TP, TDP, and SRP) relative to other shelf sites (exclusive of site 2; Figure 2, Table 6).
3. the deep water sites (6, 8 and LSC) had the lowest concentrations of total phosphorus (TP) and turbidity (T_n), on average, of the monitored sites (Figure 2, Table 6).
4. substantial spatial variations were observed within the southern end of the lake (“shelf”; exclusive of site 2) for most parameters included in the monitoring program (Figure 2, Table 6).
5. variances of measures of trophic state (Chl, TP, and T_n) were greater for the south shelf sites than for deep water sites (sites 6, 8 and LSC; Figure 2, Table 6).
6. the highest turbidity values measured in 2003 were associated with high runoff intervals during April and October (Figure 12c).
7. more than 65% of the phosphorus was in a particulate form [e.g., (TP-TDP)/TP] at all sites, on a monitored period average basis.
8. average concentrations of TP, TDP and SRP were higher in the eastern portion (sites 1 and 7), compared to other sites (4 and 5) on the shelf (Table 6).
9. chlorophyll concentrations, on a monitoring period average basis, were relatively similar across the spatial bounds of sampling, though substantial spatial variability was observed on individual days (Figure 2f, Table 6).
10. temperatures were relatively uniform over the monitored bounds of the upper waters of the lake during the period of measurements (Figure 2g, Table 6).
11. temperatures, measured hourly at the “pile cluster”, dropped precipitously on July 10 and 11, suggesting the occurrence of a seiche (Figure 2g). Similar temperature patterns were observed on two occasions in September.
12. turbidity (T_n) values and concentrations of soluble reactive phosphorus (SRP) were essentially equal in the LSC influent and effluent, with the exception of higher turbidity levels in the effluent on four occasions during the October – December interval (Figure 2j-l). The cause(s) of these higher values is unknown.

13. total phosphorus (TP) concentrations in the LSC effluent were uniformly low ($11.5 \pm 2.3 \mu\text{g}\cdot\text{L}^{-1}$) over the January – September interval of 2003 (Figure 2j). TP concentrations were higher and more variable from October to December ($17.8 \pm 3.4 \mu\text{g}\cdot\text{L}^{-1}$).
14. the concentration of total phosphorus (TP) in the LSC effluent was less than the concentration on the south shelf on most sampling days (Figure 2m); on average, the concentration was $4.9 \mu\text{g}\cdot\text{L}^{-1}$ lower (Table 8).
15. the concentration of soluble reactive phosphorus (SRP) was higher in the LSC effluent than on the shelf on most sampling days (Figure 2n), consistent with projections made in the **Draft Environmental Impact Statement** (Stearns and Wheler, 1997); on average, the concentration was $2.9 \mu\text{g}\cdot\text{L}^{-1}$ higher (Table 8).
16. turbidity (T_n) values for the LSC effluent were less than or equal to values on the shelf on most sampling days (Figure 2o); on average, turbidity was lower by 1.0 NTU (Table 8).
17. dissolved oxygen concentrations at site 3 were within 10 % of saturation (equilibrium with the atmosphere) over much of the study interval (Figure 2i). Notable exceptions were the super-saturated conditions (110-140%) during July and August and under-saturation (< 90%) in September and October.
18. beam attenuation coefficient (BAC) was generally low in the upper waters at the LSC site during the study interval of 2003 (Figure 3).
19. beam attenuation coefficient (BAC) values at the LSC site increased markedly from 40 m to the bottom on 8/14/03, following a storm on 8/10/03 (Figure 3). Modest increases in BAC were observed near the bottom of the LSC site on several monitored dates, indicating the occurrence of small increases in turbidity with the approach to the bottom at this site (Figure 3).
20. chlorophyll fluorescence profiles indicate subsurface peaks in phytoplankton concentrations occurred at the LSC intake site during the stratification period of 2003 (Figure 3). These peaks usually occurred above, or at, the maximum temperature (i.e., density) gradient, at depths ≤ 20 meters.
21. Secchi disc transparency (SD) was observed to extend beyond the lake depth at multiple sites on several occasions during the 2003 study interval (Appendix 1).
22. the 2003 results continue to support turbidity (T_n) as an alternate measure of light penetration in shallow portions of the shelf (Figure 4).

23. phosphorus loading from the Ithaca Area WWTP was lower in 2003 than in the preceding five years (Table 7, Figure 5).
24. LSC contributed ~ 4.0% of the TP load to the shelf over the May – October interval of 2003, a smaller contribution than projected (4.8%) in the **Draft Environmental Impact Statement** (Stearns and Wheler 1997; Table 7, Figure 5).
25. the Fall Creek hydrograph for 2003 was characterized by a number of major runoff events during the June – October interval (Figure 6a). Compared to long-term median values, Fall Creek flows were near normal during April, low during May, and unusually high during the June – October interval (Figure 6a-b). The median flow for the April – October interval of 2003 ranked seventh highest of the 79 years of record (1925 – 2003) for Fall Creek.
26. major wind events out of the north to northwest were not apparent during the study interval of 2002 (Figure 6c). Annual median wind speeds were somewhat lower in 2003 compared to the 1998-2002 interval (Figure 13b).
27. the 2003 results continue to support the position that TP and T_n are systematically flawed indicators of trophic state on the shelf.
28. the 2003 results continue to support the findings of Effler et al. (2002), that inorganic particles, rather than phytoplankton, are the primary regulator of T_n and SD on the shelf.
29. summer average concentrations of TP and Chl for deep water sites continue to be consistent with mesotrophy, an intermediate level of primary productivity (Tables 9 and 10).
30. study period median values for TP on the shelf were similar for 1998, 2000, 2001 and 2002, and distinctly lower for 1999 and 2003 (Figure 13d).
31. study period median values for Chl on the shelf were lowest in 1999 and highest in 1998 and 2003 (Figure 13e).
32. study period median values for T_n on the shelf were lowest for the low runoff years of 1999 and 2001 (Figure 13f).
33. no conspicuous changes in water quality have been observed on the shelf since start-up of the LSC facility in July 2000 (Figure 2; Upstate Freshwater Institute 1999, 2000, 2001, 2002, 2003).

5. Summary

This report presents the design and salient findings of a water quality monitoring study conducted for Cayuga Lake in 2003, sponsored by Cornell University. This is the sixth annual report for a monitoring program that has been conducted annually since 1998. A number of noteworthy findings are reported here for 2003 that have value for lake management. Water quality on the south shelf can vary substantially from year to year. Potential sources of variation include interannual differences in runoff, loading from WWTPs, and wind. Runoff was high during the 2003 study interval, with the median flow for Fall Creek ranking seventh highest of the 79 year record. The median velocity of winds out of the north to northwest was somewhat lower in 2003 than in the five preceding study years (1998 – 2002). On a summer (June – September) average basis, the concentration of total phosphorus was lower in 2003 than in any year of the LSC monitoring program, both on the shelf and in the main lake. On a study (April – October) average basis, total phosphorus was lowest in 1999. The summer average chlorophyll concentration on the shelf for 2003 was elevated slightly compared to the 1998 – 2002 study years; though chlorophyll concentrations remained lower than measured during the mid-1990s. The 2003 results continue to support the position (Effler et al. 2002), that inorganic particles, rather than phytoplankton, are the primary regulator of clarity on the shelf. Summer average concentrations of total phosphorus and chlorophyll **a** for deep water sites continue to be consistent with mesotrophy, a classification shared by seven of the eleven Finger Lakes. Total phosphorus concentrations and turbidity values were generally lower, and SRP (soluble reactive phosphorus) concentrations were generally higher, in the LSC effluent than on the shelf. LSC contributed ~ 4.0% of the TP load to the shelf over the May – October interval of 2003, a smaller contribution than projected in the **Draft Environmental Impact Statement**. No conspicuous changes in water quality have been observed on the shelf since start-up of the LSC facility in July 2000.

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Appendix I

Data Listing

Total Phosphorus ($\mu\text{gP}\cdot\text{L}^{-1}$)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	23.8	10.2	15.0	18.3	12.8	9.5	11.4	12.6	19.4	24.5	17.3	12.7	15.1	13.0	42.8	37.4
2	24.4	71.6	17.2	22.6	17.4	16.0	13.8	13.2	31.8	48.9	51.6	54.6	44.7	32.1	53.0	50.9
3	18.0	17.4	12.8	17.4	15.1	12.9	9.9	8.8	-	16.6	18.2	30.8	16.7	21.7	24.7	30.3
4	18.7	47.6	15.9	12.8	9.5	7.1	15.1	7.8	28.2	16.9	7.0	15.7	21.5	10.6	10.0	10.0
5	17.9	22.1	14.3	13.6	14.5	9.8	11.5	8.8	20.2	14.5	10.5	10.2	9.6	14.3	9.3	23.1
6	15.6	11.5	13.9	10.9	15.8	7.5	10.9	8.5	18.2	13.4	13.7	10.9	11.2	11.6	9.3	12.6
7	20.6	16.4	19.6	22.0	12.1	10.3	12.8	19.1	30.2	21.1	20.2	12.1	26.1	28.8	112.3	70.6
8	10.9	9.6	10.2	10.6	9.5	7.2	8.8	9.5	16.3	9.2	10.8	10.8	11.2	10.9	10.0	12.6
LSCT	12.3	11.9	12.4	10.9	7.8	8.8	9.0	7.5	17.6	14.7	7.6	10.2	10.0	13.0	10.0	11.0
LSCB	9.9	10.0	12.8	13.9	11.5	11.0	13.8	8.2	-	5.7	8.3	7.9	10.9	13.7	10.0	15.0
LSC3B	10.2	10.4	12.4	15.4	12.1	11.8	13.5	5.5	-	11.8	8.0	7.6	10.6	16.7	6.3	14.0

Total Dissolved Phosphorus ($\mu\text{gP}\cdot\text{L}^{-1}$)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	13.3	6.0	6.1	2.2	3.2	1.8	2.6	3.4	2.8	6.9	4.1	9.2	2.8	4.1	4.0	17.4
2	14.0	7.7	5.8	1.6	3.6	3.1	3.3	3.3	3.8	5.8	11.1	18.7	9.1	16.4	2.8	24.2
3	11.1	6.0	4.1	1.6	2.3	2.9	2.0	2.5	3.5	2.7	3.9	10.1	3.6	7.5	2.6	10.2
4	11.8	5.9	5.2	1.3	2.7	3.0	2.6	2.2	2.8	2.8	4.2	9.5	2.9	3.5	2.3	5.2
5	9.6	6.6	3.8	1.6	2.3	1.2	2.6	2.2	2.8	6.1	3.1	7.4	2.6	4.1	2.6	9.2
6	8.6	5.1	3.8	1.6	2.3	1.6	2.6	2.9	2.8	2.7	3.3	9.4	2.6	3.5	3.3	4.2
7	13.9	5.1	6.8	2.0	3.0	1.3	3.3	3.5	6.7	3.3	5.8	9.4	4.6	5.8	7.3	32.6
8	7.5	3.6	4.8	1.0	2.0	0.7	2.5	2.9	3.1	2.3	2.0	7.4	2.6	4.0	3.7	3.9
LSCT	7.5	5.5	4.4	1.0	2.0	1.7	4.3	2.5	3.1	2.3	2.3	8.1	2.3	3.1	3.7	4.9
LSCB	6.5	4.0	6.4	5.5	6.2	5.7	6.9	4.2	-	3.0	6.1	10.7	6.5	8.8	4.7	8.9
LSC3B	6.5	4.7	6.8	5.5	5.9	3.5	6.2	2.5	-	7.1	5.8	10.4	7.5	9.1	2.6	8.9

Soluble Reactive Phosphorus ($\mu\text{gP}\cdot\text{L}^{-1}$) values reported as 0.2 are $\frac{1}{2}$ the limit of detection ($0.3*0.5 = 0.15$) rounded to one decimal place

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	7.6	5.1	1.2	0.4	0.5	0.2	0.6	0.2	0.3	3.3	0.6	6.9	0.2	1.2	2.0	11.9
2	7.3	6.2	1.1	0.4	0.8	0.7	0.3	0.2	0.4	4.0	3.9	12.8	5.9	13.1	2.0	15.0
3	6.0	5.1	0.2	0.2	0.5	0.5	0.3	0.2	0.4	0.9	0.2	5.4	0.4	4.8	1.7	5.5
4	6.3	5.9	0.9	0.4	0.3	0.2	0.2	0.2	0.4	0.6	1.1	5.2	0.4	1.2	1.1	1.5
5	6.0	5.5	0.2	0.2	0.5	0.2	0.3	0.2	0.8	0.9	0.2	4.4	0.2	1.2	1.1	5.0
6	5.7	4.9	0.2	0.2	0.4	0.2	0.3	0.2	0.4	0.2	0.2	5.6	0.2	1.1	1.3	0.8
7	6.3	5.2	1.8	0.2	0.3	0.2	1.0	0.2	0.7	0.2	0.2	5.7	0.6	1.8	3.5	21.9
8	3.4	4.2	0.6	0.2	0.2	0.2	0.3	0.2	0.5	1.9	0.2	4.4	0.2	0.9	1.0	0.8
LSCT	4.4	4.7	0.2	0.2	0.3	0.2	0.2	0.2	0.4	0.5	0.2	5.6	0.2	0.9	1.3	1.2
LSCB	4.1	4.5	2.8	4.6	4.0	4.0	4.5	1.0	-	19.2	4.7	8.4	3.8	7.3	3.0	4.9
LSC3B	4.2	4.5	2.6	4.3	3.8	3.8	4.5	0.2	-	5.2	4.2	8.1	3.8	6.7	1.0	4.7

Chlorophyll a ($\mu\text{g}\cdot\text{L}^{-1}$)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	1.8	1.6	1.1	7.4	3.3	4.2	3.7	2.9	9.5	5.2	8.4	5.4	4.4	4.6	4.4	4.0
2	1.7	5.1	2.7	5.3	3.1	4.3	4.0	3.1	11.6	9.6	12.9	10.6	5.0	2.9	5.7	2.3
3	1.7	1.7	2.9	5.7	3.9	5.2	4.1	2.0	13.1	9.3	9.4	11.8	4.4	3.1	3.6	3.9
4	1.5	2.9	2.7	6.1	1.7	3.6	5.8	2.3	14.6	3.8	1.6	2.9	3.9	3.3	2.1	2.0
5	1.7	1.7	2.2	6.0	2.3	5.0	5.5	2.6	14.0	3.9	5.7	7.3	4.0	5.4	2.8	5.0
6	1.9	1.9	3.6	5.7	2.9	7.0	4.9	2.7	14.2	4.2	6.4	6.0	4.7	4.9	3.6	5.1
7	1.7	1.7	2.3	7.9	3.6	4.4	3.8	6.5	15.8	6.6	17.7	6.8	5.0	4.9	13.0	2.4
8	2.1	2.0	1.4	4.8	4.4	7.3	4.0	3.9	11.0	3.9	4.6	6.9	4.6	4.6	3.1	4.7
L SCT	2.0	2.1	2.3	5.2	3.0	5.5	3.8	3.1	13.2	3.0	5.8	6.9	4.9	5.9	3.5	4.0
L SCB	2.0	1.9	0.8	1.1	1.1	1.5	0.7	0.6	-	0.8	0.6	0.4	0.4	0.5	0.6	0.4
L SC3B	2.0	2.2	0.6	0.7	1.1	1.5	0.5	2.1	-	0.9	0.5	0.4	0.4	0.5	1.2	0.4

Turbidity (NTU)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	4.7	1.3	1.5	2.1	2.2	1.3	1.0	1.3	2.0	2.4	1.4	1.0	1.4	1.0	7.6	6.0
2	4.6	20.0	1.6	3.6	2.7	2.1	0.9	2.4	2.7	2.9	3.6	1.9	8.1	1.6	12.9	14.4
3	4.5	2.9	1.2	2.6	1.4	1.8	1.3	0.6	2.5	2.4	1.6	1.5	2.4	1.2	2.1	8.1
4	3.6	10.9	1.1	1.4	1.4	0.7	1.0	0.6	3.1	2.3	0.5	0.8	2.9	0.9	0.9	1.3
5	3.4	3.7	1.1	1.6	3.1	1.0	1.0	0.7	2.4	2.2	1.3	1.0	1.0	1.0	0.7	3.4
6	2.8	1.6	0.7	1.0	2.5	0.8	0.8	0.6	2.5	2.0	1.4	0.9	0.8	1.0	0.5	0.9
7	3.7	2.7	2.7	2.3	1.8	1.0	1.3	2.6	3.0	2.3	2.2	1.1	3.0	2.3	31.1	13.9
8	1.1	0.9	0.6	0.9	1.7	0.7	0.6	0.7	1.9	2.1	1.3	1.0	0.9	1.4	0.8	1.4
L SCT	1.6	1.5	0.7	1.0	1.6	0.8	0.7	0.8	2.2	1.8	1.1	0.8	0.8	1.2	0.6	0.7
L SCB	1.4	1.2	1.2	1.6	2.9	1.7	1.5	1.6	-	0.9	1.2	0.8	1.3	2.3	1.1	1.4
L SC3B	1.3	1.2	1.0	1.2	2.6	1.9	2.3	0.7	-	1.8	0.9	0.7	1.1	1.7	0.5	1.4

CaCO₃ Turbidity (NTU)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	1.1	0.2	0.5	0.3	0.4	0.4	0.1	0.2	0.3	0.9	0.0	0.1	0.4	0.1	1.9	1.3
2	0.9	9.5	0.4	0.5	0.7	0.2	0.0	0.5	0.5	0.9	1.6	0.4	1.4	0.3	5.4	2.8
3	0.7	0.8	0.5	0.6	0.3	0.2	0.7	0.0	0.4	1.3	0.6	0.3	0.6	0.2	0.0	1.5
4	0.8	3.1	0.3	0.1	0.3	0.2	0.0	0.0	0.6	1.1	0.2	0.2	0.5	0.2	0.0	0.5
5	0.8	0.9	0.2	0.2	0.7	0.3	0.1	0.1	0.4	1.1	0.5	0.7	0.3	0.1	0.0	0.7
6	0.4	0.4	0.1	0.1	0.4	0.2	0.0	0.0	0.3	1.2	0.7	0.2	0.1	0.1	0.1	0.1
7	0.4	0.4	0.9	0.4	0.2	0.3	0.3	0.3	0.4	1.0	0.4	0.1	0.7	0.9	8.0	2.5
8	0.3	0.2	0.1	0.1	0.3	0.2	0.0	0.2	0.4	1.4	0.8	0.4	-	0.5	0.2	0.7
L SCT	0.4	0.6	0.1	0.2	0.3	0.1	0.1	0.1	0.5	1.2	0.5	0.1	0.1	0.2	0.2	0.2
L SCB	0.3	0.3	0.4	0.4	0.6	0.3	0.0	0.4	-	0.4	0.5	0.4	0.5	0.9	0.4	0.6
L SC3B	0.2	0.3	0.2	0.2	0.7	0.6	0.5	0.2	-	0.6	0.3	0.2	0.3	0.3	0.1	0.5

Alkalinity (mg CaCO₃·L⁻¹)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	104.3	101.6	107.7	108.5	107.3	109.2	108.4	104.6	104.8	102.6	101.6	98.2	103.6	104.4	105.9	110.3
2	104.0	113.4	108.1	109.6	111.0	113.4	113.5	106.4	107.4	112.4	119.5	136.5	118.0	126.5	105.2	110.5
3	101.4	102.6	106.2	109.1	107.2	112.9	110.4	106.4	105.4	99.9	107.4	124.5	103.9	115.0	105.4	114.5
4	102.4	99.5	108.8	107.2	107.2	108.1	112.4	104.9	106.4	98.9	95.9	103.4	107.4	104.4	107.9	104.4
5	101.4	103.4	109.1	108.0	107.2	108.1	110.9	106.4	107.4	98.9	100.6	103.4	102.4	105.9	104.9	104.9
6	101.9	102.4	106.2	107.2	107.8	109.1	111.4	105.4	106.9	97.4	99.9	101.7	102.4	103.4	104.9	105.9
7	103.8	103.4	113.4	111.0	108.1	109.1	110.4	109.4	107.4	100.4	113.5	109.4	109.1	113.0	130.5	117.0
8	106.2	102.9	107.2	106.7	107.7	105.3	110.1	105.4	104.4	97.4	98.4	101.4	101.9	103.0	105.0	106.4
L SCT	105.3	103.4	105.3	108.1	107.2	107.2	110.4	106.4	105.4	97.4	98.4	100.9	101.4	104.4	105.4	105.9
L SCB	105.3	101.9	108.1	107.7	108.1	106.2	109.4	105.4	-	107.4	108.4	109.9	106.9	108.9	108.9	109.4
L SC3B	105.3	102.4	109.1	107.2	108.6	108.1	109.4	104.9	-	109.9	106.4	106.9	107.4	108.9	106.4	108.9

Secchi Disc (m)

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	1.5	3.8	3.0	1.5	2.0	bottom	3.9	3.3	2.1	2.4	3.0	3.5	3.0	bottom	1.3	1.7
2	1.2	0.5	bottom	1.6	2.1	3.2	bottom	1.8	1.9	1.8	2.1	1.9	1.0	bottom	0.8	0.5
3	1.6	2.8	bottom	1.9	1.6	bottom	3.6	bottom	1.9	2.5	bottom	2.6	2.6	bottom	1.8	1.6
4	2.5	0.8	3.4	2.6	bottom	bottom	3.0	bottom	2.0	2.3	bottom	bottom	2.0	bottom	bottom	bottom
5	2.0	2.1	4.0	2.2	1.3	4.5	-	5.3	1.9	2.4	3.3	3.7	5.0	bottom	5.0	3.2
6	2.6	5.0	5.0	3.1	2.0	4.5	4.1	5.1	2.0	2.5	3.6	4.1	5.3	6.9	5.0	5.9
7	1.5	1.7	2.2	1.6	2.7	bottom	bottom	2.1	1.8	2.4	2.2	bottom	1.5	bottom	0.5	0.6
8	8.0	7.0	8.0	3.0	2.4	4.5	5.5	5.5	4.0	2.3	4.3	4.5	4.0	8.0	-	7.0
L SCT	4.6	-	6.5	3.3	3.1	4.7	4.8	5.0	2.0	2.3	3.6	4.2	5.6	6.0	5.0	7.0
L SCB	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
L SC3B	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-

Temperature (°C) @ 2m

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Site:																
1	7.5	4.0	7.2	10.6	10.1	9.3	16.6	18.6	21.5	26.3	23.4	20.3	16.5	-	10.4	9.0
2	6.9	5.9	8.1	11.0	9.1	10.0	17.3	18.7	21.0	25.1	23.0	20.2	16.0	-	10.7	8.7
3	6.7	3.9	7.3	10.8	9.0	9.9	17.0	19.0	21.1	26.6	23.2	20.4	16.0	-	10.7	8.7
4	5.4	6.3	8.2	10.7	8.6	10.3	17.0	19.2	21.1	26.7	23.5	20.6	14.9	-	10.0	10.1
5	6.0	4.3	6.6	10.9	9.2	9.9	16.8	19.4	21.5	26.4	23.8	20.6	16.9	-	11.4	10.0
6	4.6	4.3	6.5	10.7	10.4	9.6	16.7	19.2	21.4	26.4	23.9	20.6	17.0	-	11.7	10.6
7	7.3	4.8	10.1	11.1	9.1	10.0	17.1	19.6	20.5	25.8	23.3	20.3	14.4	-	8.3	9.4
8	3.0	3.5	4.3	11.4	9.8	10.7	17.3	20.2	21.7	26.0	23.3	20.9	18.1	-	11.9	10.7
L SCT	3.2	4.2	5.6	10.7	9.5	10.2	16.9	19.2	22.0	26.2	23.8	20.6	16.8	-	11.7	10.5

Dissolved Oxygen (mg·L⁻¹) Site 3

Date:	4/15/03	4/24/03	5/8/03	5/29/03	6/5/03	6/19/03	7/2/03	7/17/03	7/31/03	8/14/03	8/28/03	9/11/03	9/24/03	10/9/03	10/23/03	10/29/03
Depth:																
0	12.9	12.9	12.1	10.5	10.9	-	-	-	-	10.9	9.3	10.4	9.1	11.0	10.4	9.9
1	12.5	12.8	12.0	11.5	11.5	-	-	-	-	10.4	9.3	10.6	9.1	10.9	10.2	9.6
2	12.3	12.7	12.0	12.4	11.7	-	11.9	10.8	9.8	10.9	9.4	10.6	9.0	11.0	10.2	9.7
3	12.4	12.7	12.1	12.3	11.6	-	-	-	-	10.7	9.3	10.3	8.8	11.3	10.2	9.5
4	12.5	12.7	12.2	12.4	11.6	-	-	-	-	6.9	9.2	5.5	8.2	13.0	-	9.5

Appendix 2

Lake Source Cooling Discharge Monitoring Report Data

Lake Source Cooling Discharge Monitoring Report Data

DMR Date	Temperature (Centigrade)		Flow Rate (m ³ /second)		Dissolved Oxygen (mg/L)		pH (SU)		Total Phosphorus (mg/L)		Reactive Phosphorus ^g (mg/L)	
	Daily Ave	Daily Max	Daily Ave	Daily Max	Daily Ave	Daily Max	Min	Max	Daily Ave	Daily Max	Daily Ave	Daily Max
Jul-00 ^a	10.33	10.89	1.189	1.306	11.0	11.1	7.96	8.09	0.0133	0.0136	0.005 ^b	0.005 ^b
Aug-00	10.2	11.6	1.02	1.3	11.0	11.5	8.0	8.1	0.0116	0.013	0.0059	0.0064
Sep-00	9.8	11.8	0.81	1.38	10.6	10.9	7.9	8.12	0.0122	0.0144	0.0061	0.0069
Oct-00	9.1	9.8	0.57	0.93	10.4	10.7	7.8	8.1	0.012	0.014	0.0067	0.0081
Nov-00	8.98	9.75	0.49	0.97	10.9 ^c	12.2 ^c	7.7	8.14	0.014	0.016	0.006	0.008
Dec-00	8.2	9.5	0.48	0.67	12.49	12.49	7.85	7.85	0.0109	0.0109	0.0059	0.0059
Jan-01	7.3	7.6	0.39	0.52	-	-	-	-	-	-	-	-
Feb-01	8.15	8.6	0.26	0.34	17.59	20.33	7.93	8.06	0.0095	0.011	0.0044	0.0049
Mar-01	6.56	8.67	0.31	0.44	15.76	18.18	8.0	8.1	0.0105	0.0116	0.0038	0.0042
Apr-01	7.9	9.6	0.47	0.70	15.5	17.6	7.97	8.06	0.012	0.014	0.008	0.008
May-01	9.1	10.0	0.66	0.86	15.02	18.39	7.9	8.1	0.0114	0.0139	0.0043	0.0053
Jun-01	10.4	11.4	0.97	1.31	12.01	12.34	7.96	8.08	0.0127	0.0147	0.0049	0.0058
Jul-01	10.3	11.8	0.98	1.45	11.46	11.59	7.9	8.02	0.012	0.015	0.005	0.0056
Aug-01	10.7	11.78	1.19	1.52	11.27	11.39	7.84	8.02	0.0139	0.0154	0.0062	0.0069
Sep-01	9.7	10.8	0.81	1.30	10.84	10.90	7.87	7.95	0.0141	0.0148	0.0068	0.0073
Oct-01	9.22	10.67	0.64	1.05	10.57	10.79	7.84	8.05	0.0120	0.0135	0.0049	0.0061
Nov-01	9.50	10.44	0.56	0.99	10.41	10.55	7.85	7.88	0.0122	0.0137	0.0061	0.0064
Dec-01	9.44	10.56	0.48	0.82	10.27	10.35	7.72	7.92	0.0125	0.0128	0.0060	0.0064
Jan-02	9.22	9.44	0.44	0.45	10.55	11.17	7.92	7.96	0.0104	0.0110	0.0043	0.0047
Feb-02	7.89	8.94	0.43	0.44	11.83	11.97	7.69	7.90	0.0155	0.0173	0.0049	0.0052
Mar-02	8.28	9.33	0.38	0.44	12.21	12.57	7.83	7.90	0.0121	0.0161	0.0038	0.0043
Apr-02 ^f	9.11	10.94	0.53	1.06	11.69	11.88	7.92	7.98	0.0178	0.0323	0.0037	0.0042
May-02	9.72	10.78	0.68	1.13	11.53	11.75	7.77	8.02	0.0108	0.0116	0.0029	0.0044
Jun-02	10.67	11.83	1.09	1.33	11.08	11.26	7.89	8.06	0.0108	0.0121	0.0039	0.0042
Jul-02	10.72	12.00	1.47	1.92	11.30	12.79	7.75	7.89	0.0142	0.0178	0.0042	0.0056
Aug-02	10.50	11.50	1.41	1.82	12.84	15.58	7.75	7.93	0.0095	0.0103	0.0038	0.0047
Sep-02	10.00	11.00	1.2	1.8	15.21	20.85	8.0	8.0	0.0096	0.0110	0.0037	0.0047
Oct-02	9.4	10.3	0.7	1.8	12.73	24.68	7.8	8.1	0.0118	0.0136	0.0056	0.0066
Nov-02	9.2	10.3	0.6	1.7	9.96	10.40	7.6	8.0	0.0122	0.0139	0.0062	0.0065
Dec-02	8.6	9.1	0.6	1.2	10.54	10.79	7.5	8.1	0.0083	0.0100	0.0033	0.0040

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DMR Date	Temperature (Centigrade)		Flow Rate (m ³ /second)		Dissolved Oxygen (mg/L)		pH (SU)		Total Phosphorus (mg/L)		Reactive Phosphorus ^a (mg/L)	
	Daily Ave	Daily Max	Daily Ave	Daily Max	Daily Ave	Daily Max	Min	Max	Daily Ave	Daily Max	Daily Ave	Daily Max
Jan-03	8.2	9.2	0.4	0.5	10.64	11.59	7.5	7.7	0.0103	0.0115	0.0037	0.0048
Feb-03	7.8	8.2	0.3	0.3	13.40	13.84	7.8	7.9	0.0095	0.0099	0.0039	0.0044
Mar-03	7.6	9.2	0.3	0.4	12.52	13.00	7.5	7.9	0.0111	0.0155	0.0032	0.0039
Apr-03	8.2	9.4	0.4	0.8	12.75	13.30	7.6	7.9	0.0138	0.0169	0.0045	0.0049
May-03	8.7	9.6	0.6	0.9	12.73	14.60	7.5	7.8	0.0120	0.0131	0.0039	0.0046
Jun-03	9.4	10.6	1.0	1.5	12.05	12.20	7.7	7.9	0.0136	0.0159	0.0038	0.0042
Jul-03	10.4	10.8	1.2	1.6	11.77	12.86	7.6	7.8	0.0111	0.0125	0.0039	0.0051
Aug-03	10.5	11.6	1.6	2.0	11.63	12.40	7.1	7.8	0.0090	0.0093	0.0051	0.0055
Sep-03	9.6	10.6	1.2	1.8	11.09	11.31	7.4	7.7	0.0128	0.0170	0.0062	0.0073
Oct-03	9.1	10.1	0.6	0.9	10.27	10.50	7.6	7.7	0.0166	0.0209	0.0065	0.0070
Nov-03	8.9	9.9	0.6	1.2	10.42	10.61	7.7	7.8	0.0201	0.0252	0.0055	0.0061
Dec-03	8.2	8.8	0.6	1.0	10.61	10.64	7.6	7.9	0.0170	0.0202	0.0048	0.0064

Lake Source Cooling Discharge Monitoring Report Data

DMR Notes:

1. To comply with changes in the NYS DEC DMR Manual for Completing the Discharge Monitoring Report for the State Pollutant Discharge Elimination System, sample measurements will be reported in the same number of significant digits that are specified in the permit. All calculations will be performed prior to any rounding, and, when rounding, if the digit being dropped is 0-4, the preceding number will be left as is, if the digit being dropped is 5-9, the preceding number will be increased. This change took effect for the reporting of September 2002.
2. Since June 2002, reactive phosphorus results below the limit of detection of 0.3 µg/L have been changed to 0.3 µg/L for all DMR calculations. Prior to this a value of ½ the limit of detection was used for DMR calculations.

^a During the month of July 2000, the Lake Source Cooling Heat Exchange Facility was commercially operational (following a brief commissioning period) from July 17 through July 31, therefore the data reported in the DMR is reflective of the 15 days of operation out of the 31 total days in the month.

^b The data reported for soluble reactive phosphorus in July 2000 is from one sampling date, 7/27/2000, during the last calendar week of July. The SPDES permit requires soluble reactive phosphorus samples to be analyzed weekly. Although a sample was collected by Cornell University during the third calendar week of July, the sample was not analyzed due to laboratory error. This error has been corrected.

^c One of the five samples analyzed for dissolved oxygen had a false high result and was eliminated from reporting on this DMR on the recommendation of our consultant/analytical laboratory, Upstate Freshwater Institute Inc.

^d The LSC discharge was shut down for emergency repairs on December 8, 2000 and remained off line for the rest of the month of December. The data reported on the DMR is reflective of monitoring conducted between December 1 and December 8 (samples collected weekly, so the data is from one sampling event).

^e Please note that there are no data presented in the DMR for effluent parameters DO, pH, total phosphorus, and reactive phosphorus. The LSC discharge was shut down for emergency repairs on December 8, 2000 and remained off line until January 29, 2001. Effluent sampling was conducted the week of January 29 as required by the permit; the effluent sample was collected on Thursday February 1. The effluent data for the sample collected during the last week of January will be included with the data presented in the February DMR.

^f Analytical results from 4/18/02 were not included in these calculations because holding times were exceeded.

^g Flow and temperature data for 6/11/03 – 6/14/03 were missing and could not be included in the calculation.